

## Design, modelling and implementation of stormwater source control technologies

### A workshop of the International Working Group on Source Control and Stormwater Management

(IWA/IAHR Joint Committee on Urban Drainage)

Workshop Organisers: Sylvie Barraud<sup>1</sup>, Gilles Rivard<sup>2</sup> & Tim Fletcher<sup>3</sup> (<sup>1</sup>Lyon 1 University /INSA Lyon, France, <sup>2</sup>Aquapraxis, Canada & <sup>3</sup>Monash University, Austalia)



Sunday June 27<sup>th</sup>, 2010

**Acknowledgement:** Thanks to the team of GRAIE (Elodie Brelot, Lucie Dupouy, Ophélie Chassard and Christiane Alonso for their assistance in coordinating and preparation of workshop materials

# Design, modelling and implementation of stormwater source control technologies

Conception, modélisation et retours d'expérience dans le domaine de la gestion des eaux pluviales à la source

#### Workshop short presentation:

The aim of the workshop was to examine aspects relating to the design and implementation of stormwater source control technologies. The workshop, following the previous SOCOMA workshop held in Lyon in 2007 (see <a href="http://graie.org/SOCOMA">http://graie.org/SOCOMA</a>), has been specifically focussed on the modelling and performance evaluation of source control techniques. It has also evaluated the lessons of implementation, based on a number of important case studies. Each of these themes has included a general presentation of the current state of the art in terms of understanding, research, before exploring particular aspects in detail, through interactive presentations and discussions.

The workshop will result in publication of review articles, summarising the current state of the art in the modelling, performance evaluation and implementation of source control strategies.

#### Présentation du Workshop :

L'objectif du workshop a été d'aborder différents aspects concourant à la conception, la mise en œuvre de stratégies de contrôle à la source en matière de gestion des eaux pluviales.

Ce workshop qui fait suite à celui de 2007 qui s'est tenu à Lyon (<u>http://graie.org/SOCOMA</u>) a été plus particulièrement consacré à la modélisation de ces systèmes et leur utilisation en terme opérationnel (pour planifier, concevoir et gérer), à l'analyse de performance et aux retours d'expérience tant en terme technique, environnemental, socio-économique qu'en terme de gouvernance.

Ces différents thèmes ont fait l'objet d'une présentation générale permettant de faire un point sur les connaissances, les recherches et les savoir faire existants sur le thème, suivie de plusieurs focus sur des aspects particuliers du thème.

Ce workshop s'est donné pour objectif de servir de support à la rédaction d'articles de synthèse internationale sur la conception, la modélisation et les problèmes d'implantation de gestion des eaux pluviales à la source.

### WORKSHOP TIMETABLE

Time slot	Workshop Activity / Topic	Presenters
9:00 am - 9:10 am	Introduction to workshop	<b>G. Rivard,</b> Chairman of SOCOMA work group, Canada
Modelling		
9:10 am - 9:20 am	General literature overview on modelling followed by a focus on :	Coordinator: S. Barraud*
9:20 am -10:00 am	Modelling urban stormwater impact mitigation by using BMPs at the catchment scale - Implementation of Source Control systems in Italy	<b>G. Freni &amp; G. Mannina</b> , Palermo University, Italy
10:00 am - 10:40 am	Opportunities and drawbacks of simulating infiltration processes	<b>S. Fach,</b> Innsbruck University, Austria
10:40 am - 11:00 am	Coffee break	
Source control perform	mance	
11:00 am - 11:10 am	General literature overview followed by a focus on :	Coordinator: <b>T. Fletcher</b> *
11:10 am-11:50 am	"Green" technologies and infrastructures for the control and treatment of impervious surface runoff	<b>B. Ellis</b> , Urban Pollution Research Centre, Middlesex University, UK
11:50 am-12:30 am	Elements in favour of source control: The experience of the French on-site observatory OTHU	<b>S. Barraud</b> , Lyon 1 University / INSA Lyon & <b>A. Foulquier</b> , Lyon 1 University, France
12:30 pm - 2 :00 pm	Lunch	
Implementation and a	doption: success, failure and lessons lear	nt
2:00pm - 2:10pm	Introduction followed by focus on :	Coordinator: G. Rivard*
2 00 pm - 2:40 pm	Lessons from a catchment-scale public & private-land retrofit project	<b>T. Fletcher &amp; M. Burns,</b> Monash University, Australia
2:40 pm - 3:20 pm	Lessons from the Shepherd Creek experiment.	<b>B. Shuster,</b> National Risk Management Research Laboratory, Office of Research and Development, USEPA, USA
3:20 pm - 3:40 pm	Coffee break	
3:40 pm - 4:00 pm	Some lessons learnt about Source Control strategies in France	<b>C. Carré</b> - Paris 1 University, France
4:00 pm – 4:40 pm	Lessons learnt and experiences about Source Control strategies in Brasil	N. Nascimento – UFMG, Brazil
4:40 pm - 5:00 pm * With the help of SOCC	Discussion	

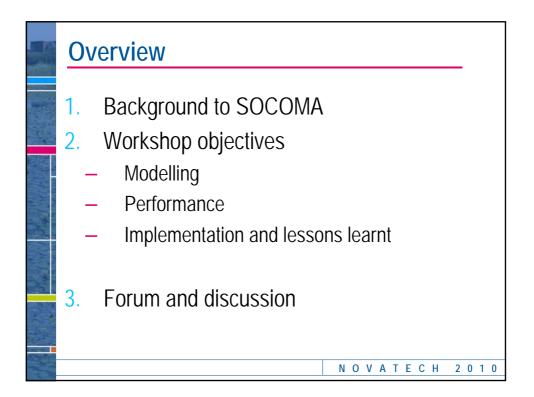
\* With the help of SOCOMA Work Group

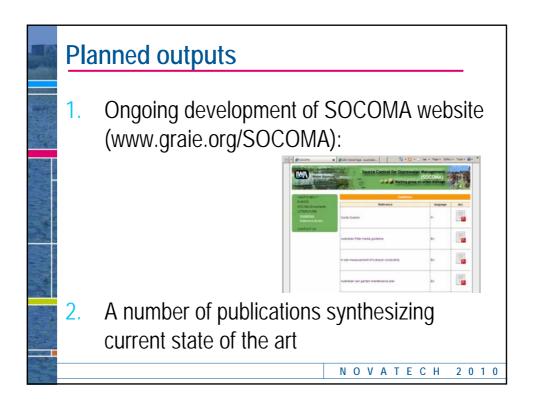
## **INTRODUCTION TO WORKSHOP**

G. Rivard

Chairman of SOCOMA work group, Aquapraxis, Canada

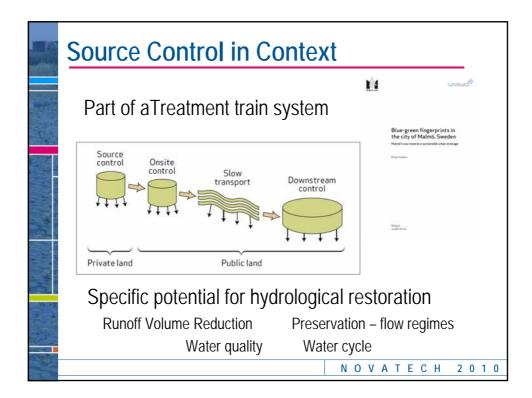


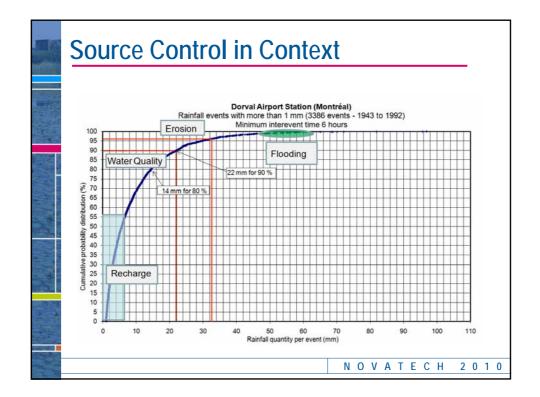




Time slot 9.00 am – 9.10 am	Workshop Activity / Topic Introduction to workshop	Presenters G. Rivard, Chairman of SOCO work group,	
Modelling			
9.10 am – 9.20 am	General literature overview on modelling followed by a focus on :	<u>Coordinator:</u> S. Barraud* G. Freni & G. Mannina, Palermo University, Italy S. Fach, Innsbruck University, Austria	
9.20 am – 10.00 am	Modelling urban stormwater impact mitigation by using BMPs at the catchment scale - Implementation of Source Control systems in		
10.00 am - 10.40 am	Opportunities and drawbacks of simulating infiltration processes		
10:40 am - 11:00 am	Coffee break		
Source control perform	mance		
11:00 am - 11:10 am	General literature overview followed by a focus on :	<u>Coordinator:</u> <b>T. Fletcher</b> *	
11:10 am-11:50 am	"Green" technologies and infrastructures for the control and treatment of impervious surface runoff	B. Ellis, Urban Pollution Researcentre, UK	
11:50 am-12:30 am	D am-12:30 am Elements in favour of source control: The experience of the French on-site observatory OTHU S. Barraud, Lyor Lyon & A. Foulquier, Lyor France		
	Lunch		

Time slot	Workshop Activity / Topic	Presenters	
	adoption: success, failure and lessons learnt		
2 00 pm – 2 10 pm	Introduction followed by focus on :	Coordinator: G. Rivard*	
2 10 pm – 2 40 pm	Lessons from a catchment-scale public & private- land retrofit project	T. Fletcher & M. Burns Monash University, Australia	
2:40 pm - 3:20 pm	Lessons from the Shepherd Creek experiment.	B. Shuster, National Risk	
		Management Research Laborato	
		Office of Research and Development,	
		Bereiopinient,	
3:20 pm - 3:40 pm	Coffee break		
3:40 pm - 4:00 pm	Some lessons learnt about Source Control strategies in	C. Carré - Paris 1 University, Fra	
4:00 pm - 4:40 pm	Lessons learnt and experiences about Source	N. Nascimento – UFMG, Brazil	
	Control strategies in Brasil		
4:40 pm - 5:00 pm			
	Discussion		





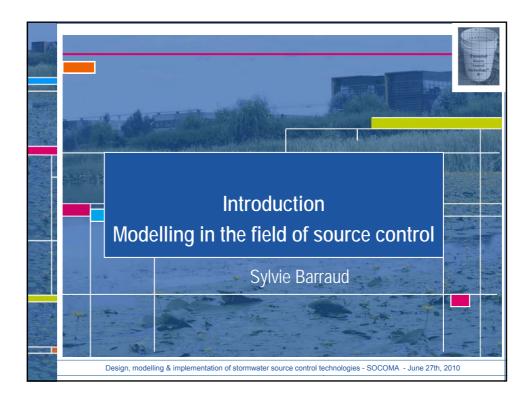
	Characteristic of source-control technologies	Ref./Rationale
Advantages	Easier implementation: smaller volumes are often technically easier to manage (for example, groundwater mounding with infiltration is less probable)     Greater private investment: cost of management system relying on source-control will be in part supported by private sector. Techniques at the parcel scale do not involve any land opportunity cost.     Greater direct private benefits (e.g. reuse of water)     Potential microclimate benefits (reduce heat island	<ul> <li>Typical design of small systems in local guidelines (e.g. Melbourne Water)</li> <li>(Environmental Services Division, 2009)</li> <li>(Brown, 2010)</li> <li>(Endreny, 2008)</li> </ul>
Drawbacks	effect)           - Limited volume treated (integration to landscape can be restrictive in dense urban context): effects of peak flows (floodings, erosion) are thus hardly mitigated           - Complexity of negotiation: implementation on private parcels is subject to public commitment; drivers for a large scale implementation may be complex.           - Few economies of scale (for construction and potentially for maintenance): larger systems generally show a lower cost per unit volume treated.           - Uncertain maintenance regimes (in private properties)	- (Burns, et al., 2010b) - (Fletcher, et al., 2010a) - (Environmental Services Division, 2009) (Wossink & Hunt, 2005) - (Environmental Services Division, 2009, Chap.7)

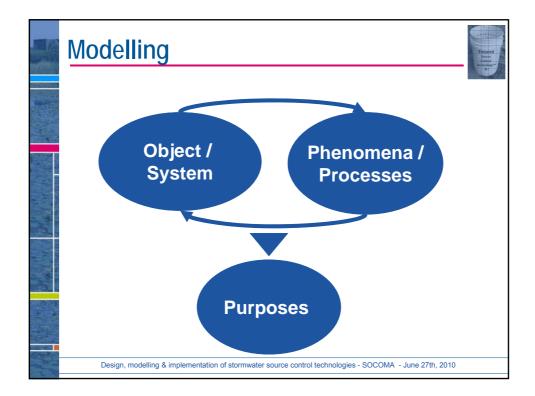
**General literature overview** 

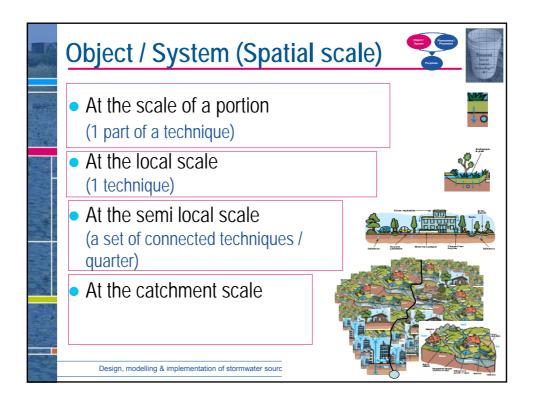
### INTRODUCTION - MODELLING IN THE FIELD OF SOURCE CONTROL

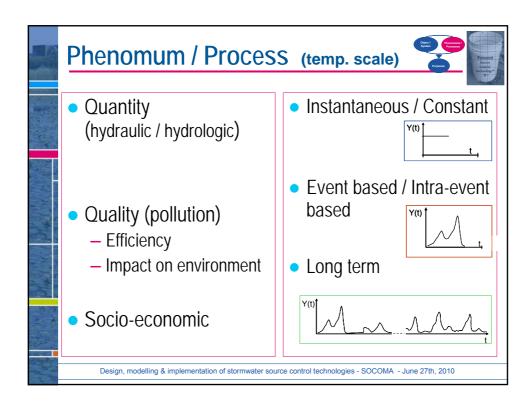
S. Barraud

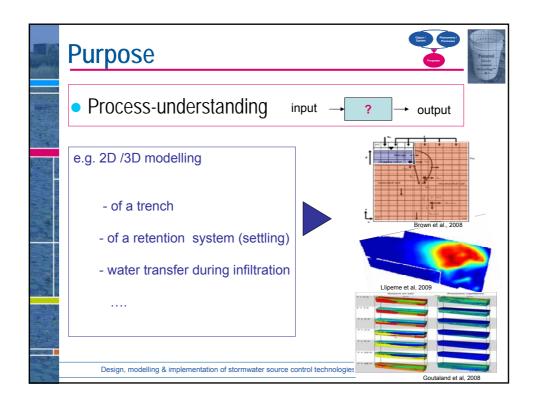
Lyon 1 University / INSA Iyon, France



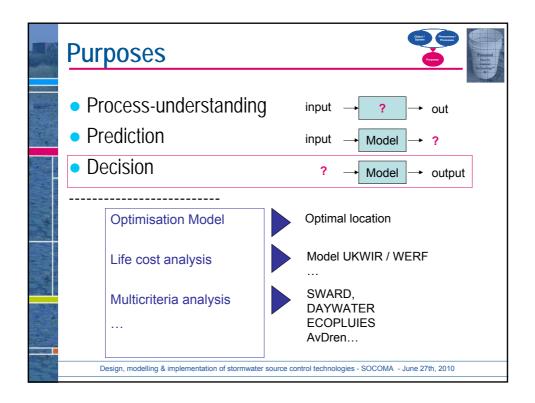


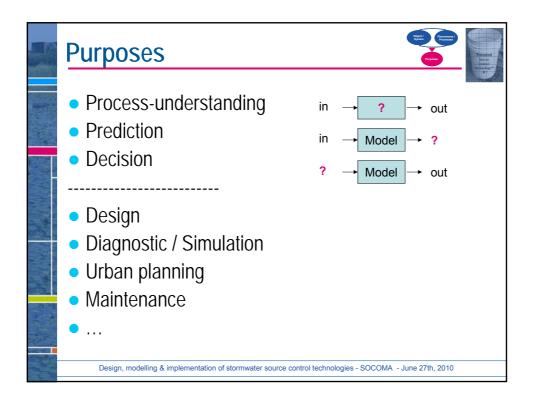


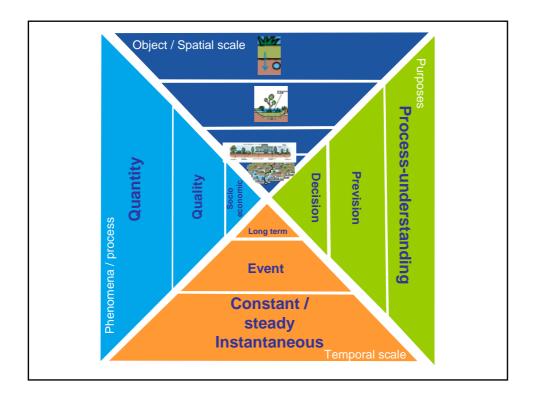


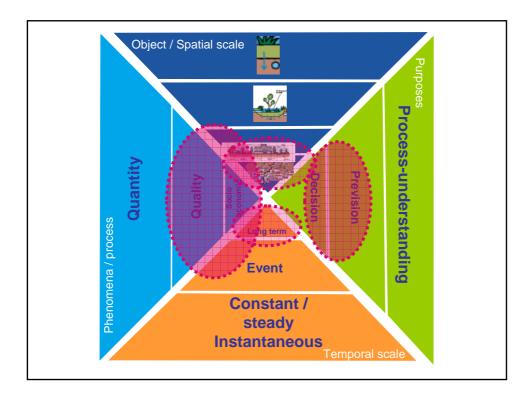


Purposes		
Process-under     Prediction	rstanding input → ? - input → Model -	→ output → ?
Quantity (run-off formation / hydr. Efficiency / impact on receiving waters)	<ul> <li>e.g.</li> <li>PULS model</li> <li>Run-off coefficient per type of structure / area</li> <li></li> <li>Hydraulic eff. Series of mixed</li> </ul>	Canoe Mouse SWMM MUSIC
Pollution (generation / treatment / impact)	<ul> <li>waterbodies /</li> <li>First order kinetic (K-C*) decay algorithm</li> <li></li> </ul>	Storm.BMP / Storm.xxl 









### MODELLING URBAN STORMWATER IMPACT MITIGATION BY USING BMPS AT THE CATCHMENT SCALE

## IMPLEMENTATION OF SOURCE CONTROL SYSTEMS IN ITALY

G. FRENI & G. Mannina

Palermo University, Italy

# Modelling

#### Modelling urban stormwater impact mitigation by using BMPs at the catchment scale. Implementation of Source Control systems in Italy

Gabriele Freni<sup>1</sup>, Giorgio Mannina<sup>2</sup>

 <sup>1</sup> Facoltà di Ingegneria ed Architettura, Università "Kore" di Enna, Cittadella Universitaria, 94100 Enna, Italy (E-mail: gabriele.freni@unikore.it)
 <sup>2</sup> Dipartimento di Ingegneria Idraulica ed Applicazioni Ambientali, Università di Palermo, Viale delle Scienze, 90128 Palermo, Italy. (E-mail: mannina@idra.unipa.it)

#### ABSTRACT

The continuous growth of urban areas and the increasing public awareness about environmental impact of stormwater raised high interest on quality impact on the receiving water quality. Indeed, in the last decades, large efforts have been provided for improving urban drainage systems in order to mitigate environmental impacts. In the last years the limitation linked to the traditional urban drainage scheme were pointed out and new approaches are developing introducing more natural methods for retaining and/or disposing of stormwater (Emerson et al. 2005; Bledsoe 2002; Niemczynowicz 1994). These mitigation measure are generally called Best Management Practices (BMP) or Sustainable Urban Drainage System or Low Impact Developments and they include practices such as infiltration and storage tanks in order to reduce the peak flow, increasing the time it takes to reach the receiving water system and retaining at least part of the polluting components. The selection of the best mitigation measure for a specific site is still a controversial topic and several factors should be taken into account. The integration of such mitigation measures in an integrated urban drainage model can provide an interesting tool for comparing different mitigation solution at catchment scale and for selecting the best technique.

In the present study, a comprehensive BMP modelling approach will be presented in order to allow one to evaluate the mitigation efficiency of different BMP schemes including infiltration and storage facilities. The comprehensive BMP modelling approach consists in two main sub-modules: the urban drainage model and the BMP model. The former enables one to assess the hydrograph and the pollutograph at the BMP inlet, on the other hand, the latter models the BMP evaluating the main processes that control the BMP outflow (in terms of both quantity and quality aspects). The urban drainage model is based on a conceptual simplified model developed during previous studies (Mannina and Viviani, 2010) and reproduces the physical phenomena that take place both in the catchments and in the sewers, allowing to determine the hydrograph and pollutograph in the sewer. For the assessment of the latter, particular care is taken about sediment transformation in sewers. considering their cohesive-like behaviour caused by organic substances and by physicalchemical changes during the sewer sediment transport. The catchment urban drainage system is modelled coupling two reservoirs in series and the phenomena that take place during both dry and wet period are developed. Indeed water quality of storm water runoff varies widely depending on the surface use and pollution dry weather. For this reason,

particular care was addressed towards the antecedent dry weather period responsible of the pollution of storm water and the first flush phenomena. Regarding the BMP model, a conceptual model taking particular care in simulating clogging phenomena that take part during their life cycle reducing mitigation efficiency was employed (Freni et al., 2009). The model introduces the concept of an "effective area" as the horizontal area below the trench bottom where the infiltration paths become linear and parallel, so the phenomenon can be considered one-dimensional. According to this definition, it is possible to use a one-dimensional model in order to estimate the infiltration process around the BMP structure and assuming equilibrium between the stored water volume in the structure and the infiltrated volume in the soil where the flow paths are vertical. The model simulates the hydraulics of an infiltration structure that is supposed to operate as a nonlinear reservoir, equipped with a weir that simulates the overflows to the drainage system or the catchment surface when the infiltration device reaches saturation. The infiltration flow is evaluated using the Green-Ampt equation.

In order to gain insight on the best technique, this study compares different distributed and centralized urban stormwater management techniques, including infiltration and storage facilities. A long-term simulation is employed to account for the effects of sediments in BMPs, which generally reduce the hydraulic capacity. The results allow us to draw some conclusions on the peculiarities of BMP techniques, on the possibility of integrating different techniques for improving efficiency and on BMP maintenance planning.

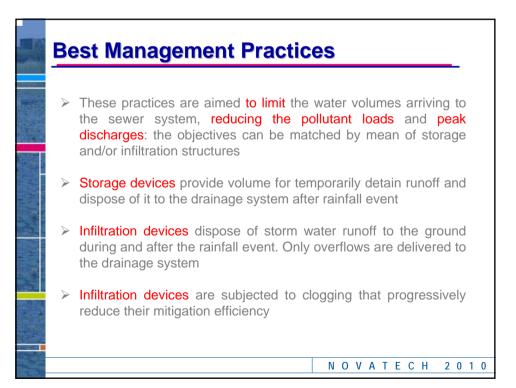
A specific survey was carried out to characterize the soils' infiltration capacities. Using the results from the case study, some general conclusions can be drawn:

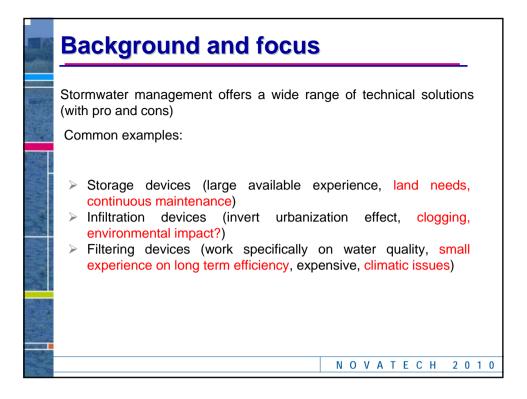
- centralized techniques are more robust and can be effective also with small specific design volumes;
- BMPs based on stormwater infiltration process can be effective if the soil infiltration capacity allows their use, but their efficiency can be reduced by clogging (in the presented case study, small infiltration structures were 40% clogged after only 6 years of service);
- mixed configurations, involving both source controls and centralized techniques are, in some cases, more efficient than centralized controls (maintaining the same design specific volume) by avoiding frequent sewer flushing during wet periods and protecting receiving waters from frequent CSO spills.

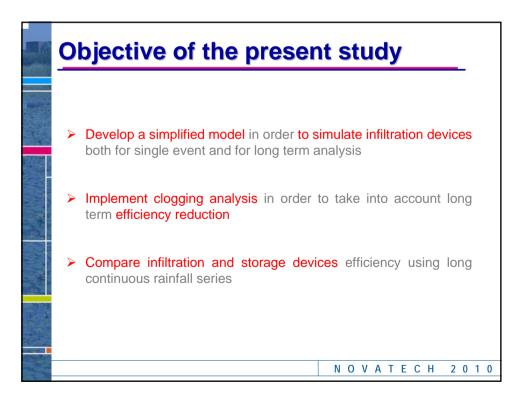
#### REFERENCES

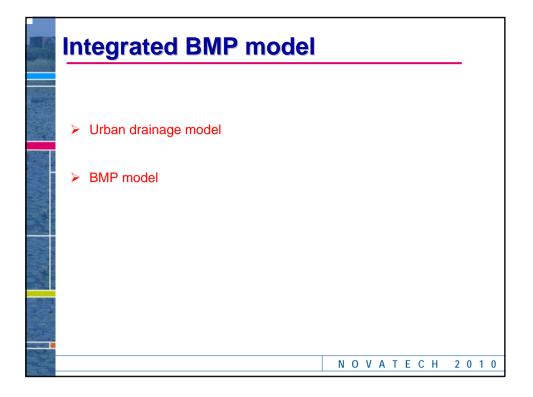
- Bledsoe, B. (2002). Stream Erosion Potential and Stormwater Management Strategies. J. Water Resour. Plng. and Mgmt. 128(6), 451.
- Emerson, C. H., Welty, C., Traver, R.G. (2005). Watershed-Scale Evaluation of a System of Storm Water Detention Basino. J. Hydrol. Eng. ASCE, 10(3), 237-242.
- Freni, G., Mannina, G. and Viviani, G. (2010) Urban stormwater quality management: centralized versus source control. J. of Water Resources Planning and Management – Asce. Volume 136(2), 268-278.
- Freni G., Mannina G. Viviani G. (2009). Stormwater infiltration trenches: a conceptual modelling approach. Water Science and Technology, vol. 60 (1) pp. 185-199.
- Mannina, G. & Viviani, G. (2010) An urban drainage stormwater quality model: model development and uncertainty quantification. Journal of Hydrology, 381(3-4), 248-265.
- Niemczynowicz J. (1994) New aspects of urban drainage and pollution reduction towards sustainability. Water Science and Technology, 30(5), 269–277.

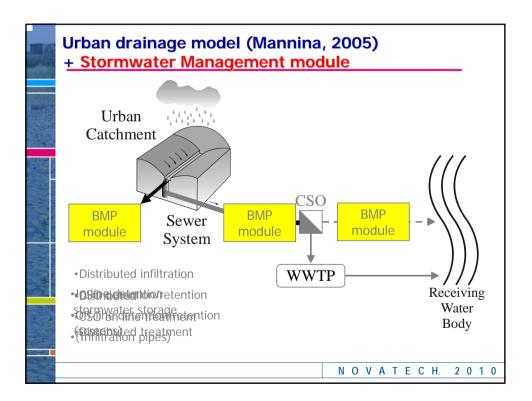


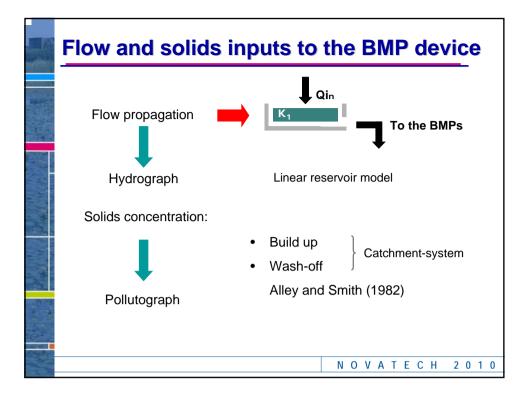


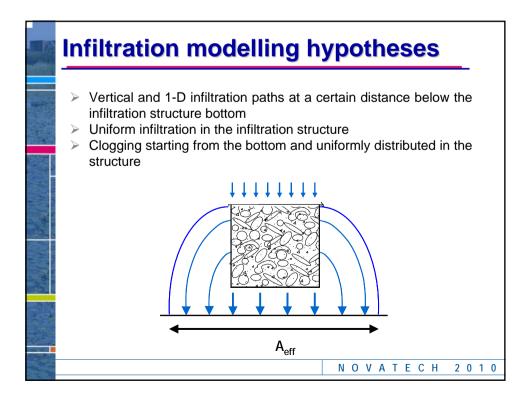


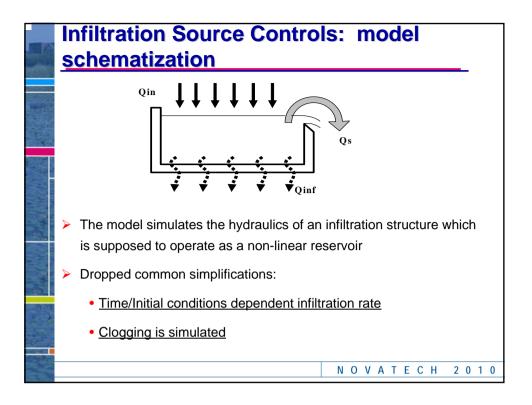


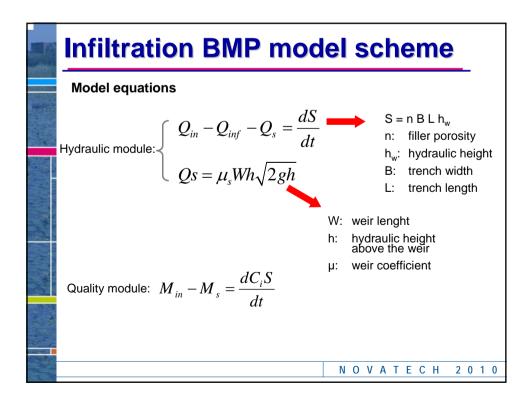


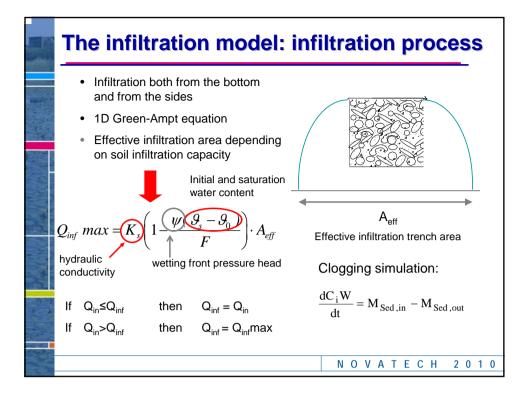


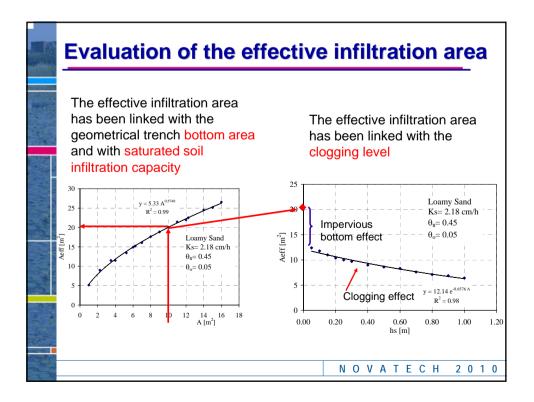


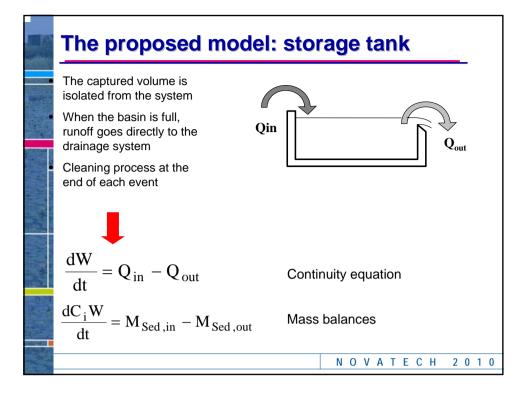


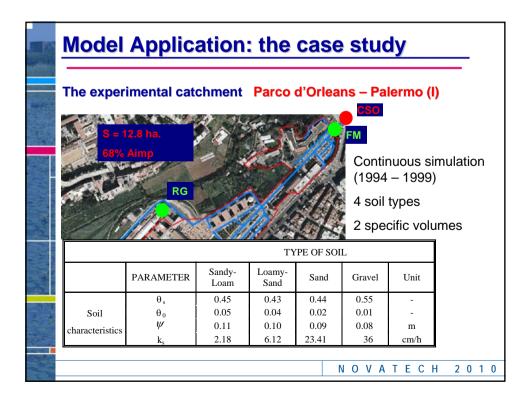










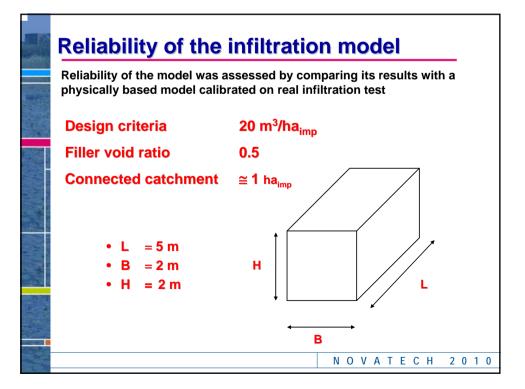


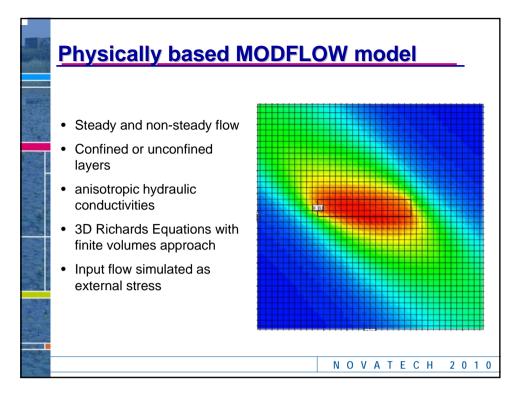
### **Model Application: rainfall data** Rainfall data are collected since 1993 with a tipping bucket raingauge and data logger at maximum time resolution of 1 sec.

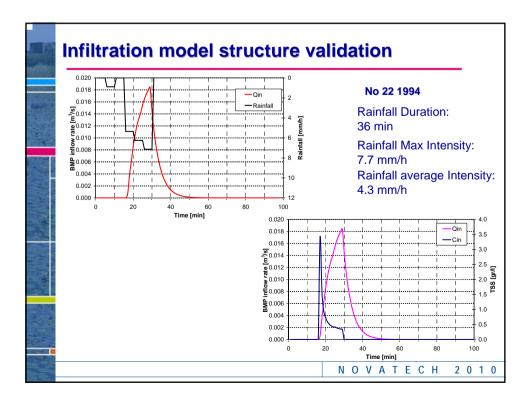
Discharge data are collected since the same year with an ultrasonic flow meter installed at basin outlet

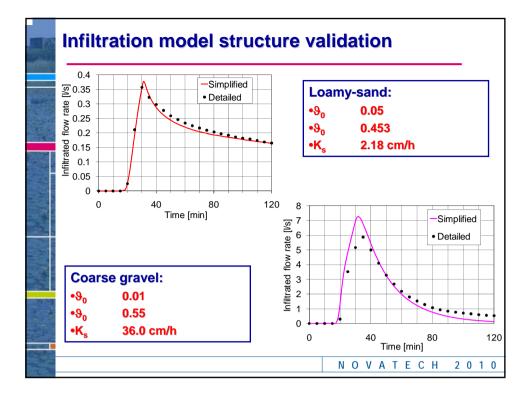
From this archive, a 6-years continuous rainfall series have been extracted and used for the simulations

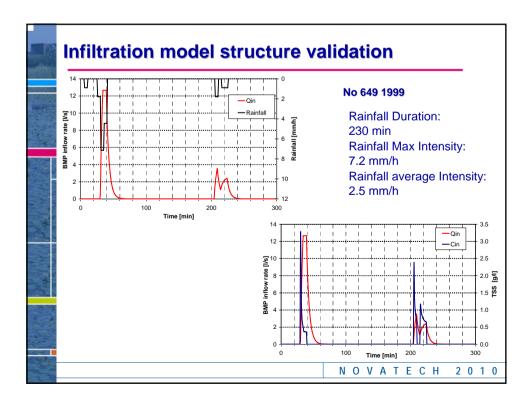
285 22 5.5	552 56 4.5	655 63	602 73	634 66	582
			73	66	
5.5	4.5	2.0		00	57
		3.8	4.3	4.1	4.6
7.2	8.5	9.7	7.7	5.8	6.2
37.8	42.2	57.8	36.5	40.2	42.8
27.3	28.5	34.3	22.4	33.6	29.2
22.1	23.2	25.6	19.8	22.7	24.2
		NO	VAT	ЕСН	20
	37.8	37.8         42.2           27.3         28.5	37.8         42.2         57.8           27.3         28.5         34.3           22.1         23.2         25.6	37.8         42.2         57.8         36.5           27.3         28.5         34.3         22.4           22.1         23.2         25.6         19.8	37.8         42.2         57.8         36.5         40.2           27.3         28.5         34.3         22.4         33.6           22.1         23.2         25.6         19.8         22.7

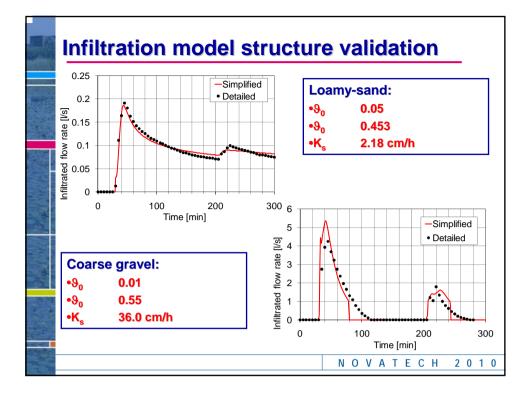


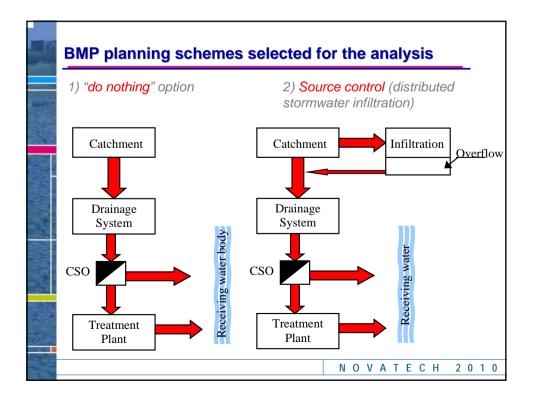


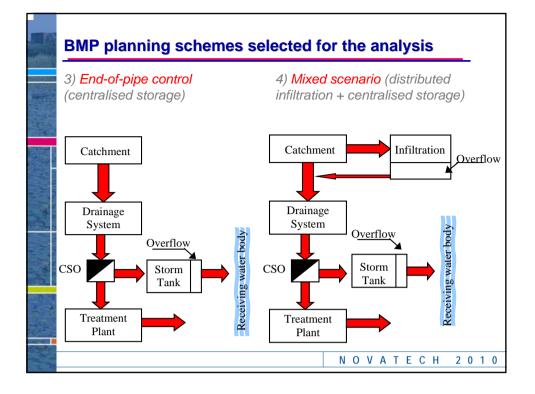


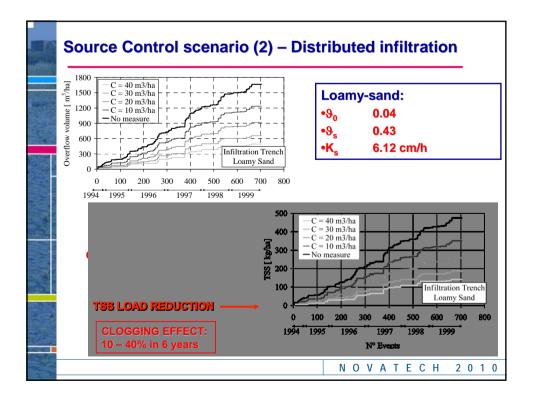


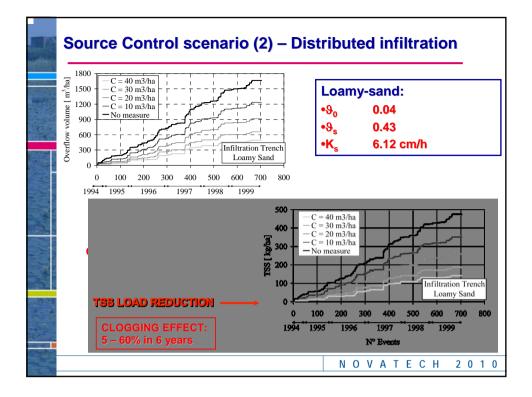


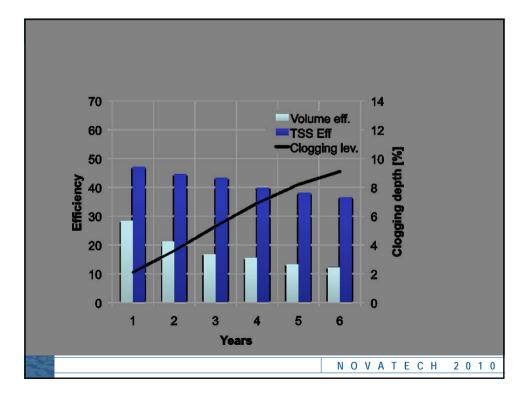


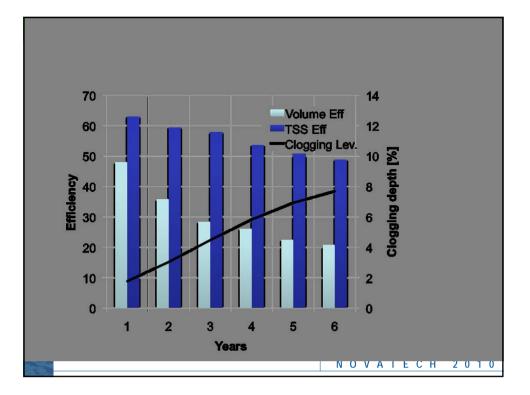


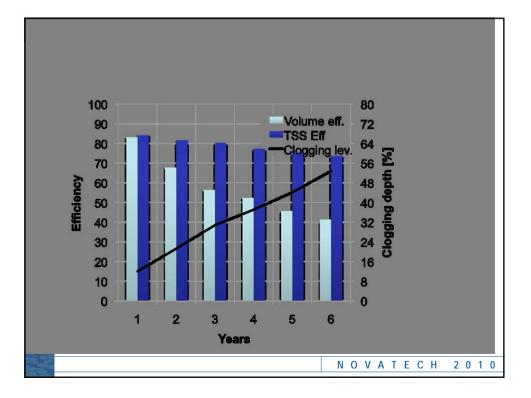


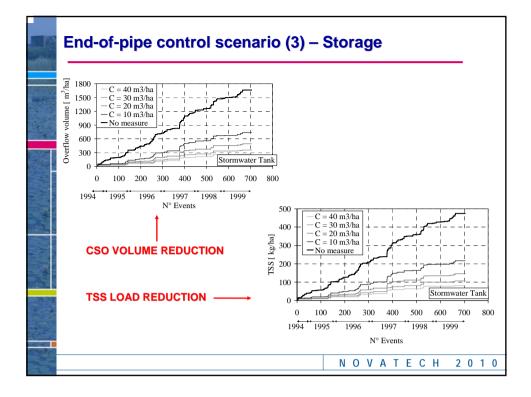


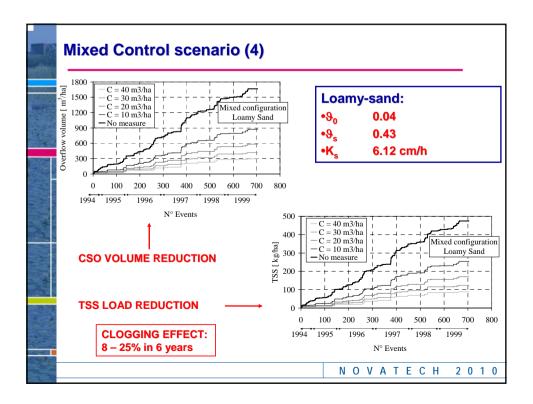


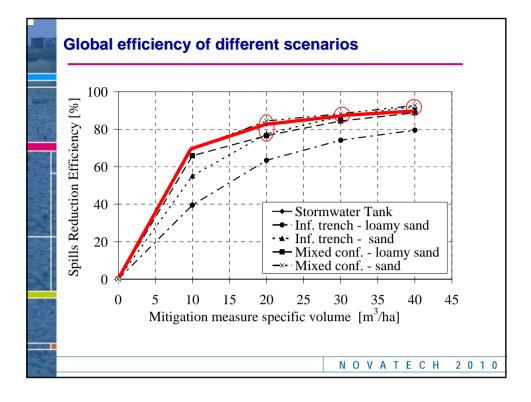


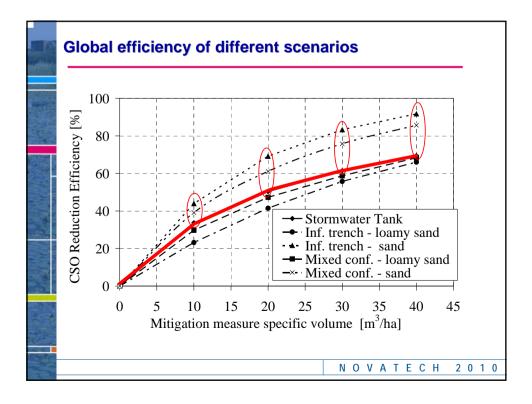


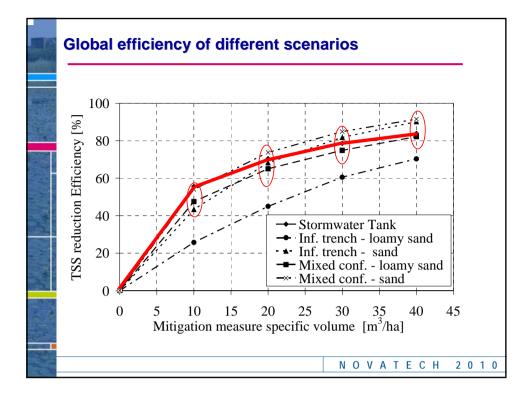


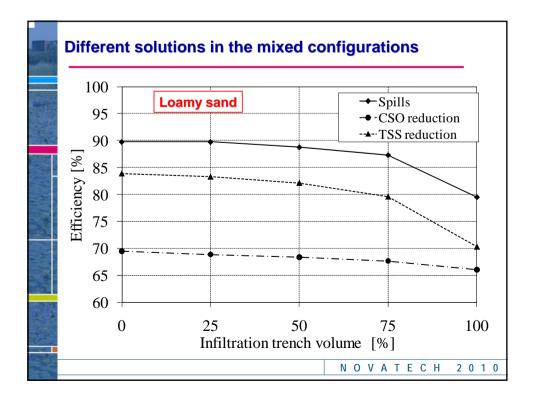


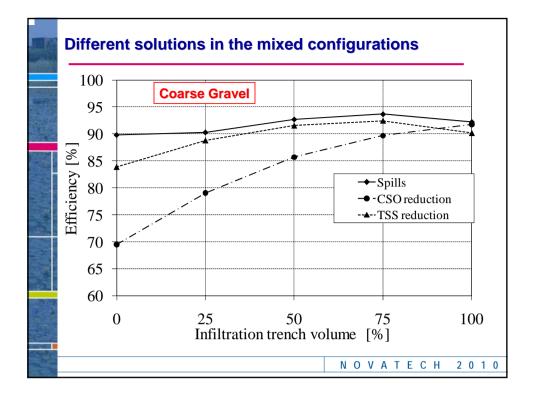


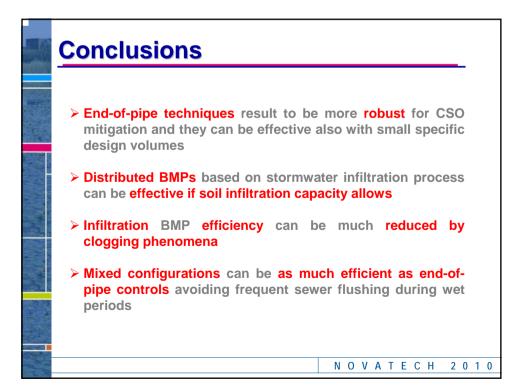








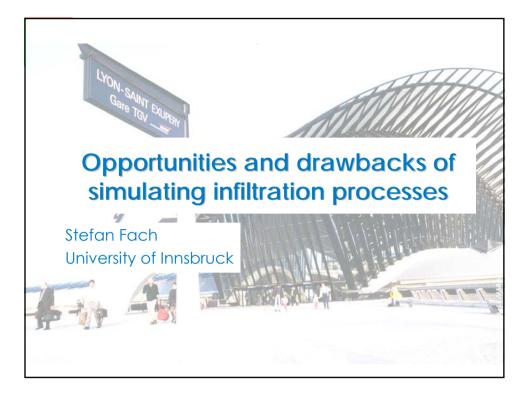


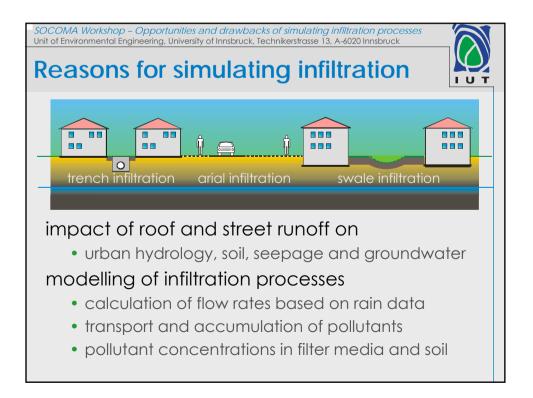


### OPPORTUNITIES AND DRAWBACKS OF SIMULATING INFILTRATION PROCESSES

S. Fach

Innsbruck University, Austria



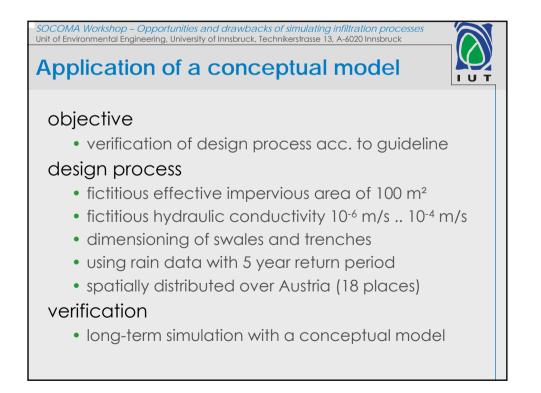


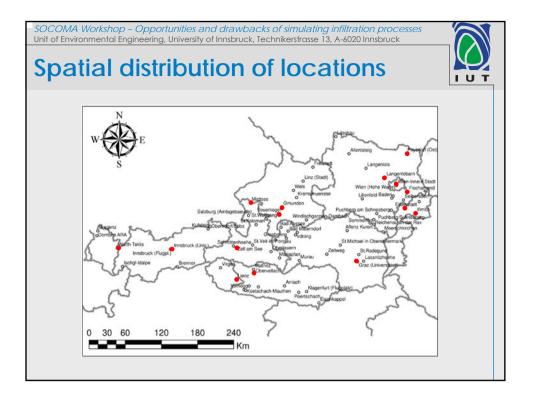
SOCOMA Workshop – Opportunities and drawbacks of simulating infiltration processes Unit of Environmental Engineering, University of Innsbruck, Technikerstrasse 13, A-6020 Innsbruck

## Comparison of infiltration models



	conceptual models	finite element models
computational effort	low (within minutes)	high (within hours)
flow process	Darcy's law	<b>Richard's equation</b>
constraints	(saturated) homogeneous soil	mesh resolution (const. characteristics per cell)
pollutant retention	-	adsorption isotherms
constraints	-	equilibrium conditions (residence time)
purpose	system analysis, i.e. interaction with other compartments in urban drainage systems	detailed understanding of flow and pollutant transport processes





	SOCOMA Workshop – Opportunities and drawbacks of simulating infiltration processes
	Unit of Environmental Engineering, University of Innsbruck, Technikerstrasse 13, A-6020 Innsbruck
1	

## Pluviographs in detail



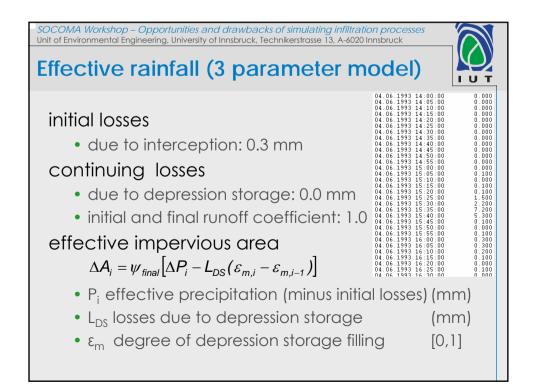
station	begin	end	location	state	evaporation
7704	01.01.1992	31.12.2006	Eisenstadt	Burgenland	625-650
7818	01.01.1991	31.12.2003	Illmitz	Burgenland	625-650
18121	01.01.1987	31.12.2006	Obervellach	Carinthia	600-625
19821	01.01.1983	31.12.2006	Weißensee-Gatschach	Carinthia	500-550
6050	01.01.1985	31.12.2003	Fischamend	Lower Austria	625-650
4081	01.01.1993	31.12.2006	Langenlebarn	Lower Austria	600-625
2503	01.01.1993	31.12.2006	Poysdorf-Ost	Lower Austria	600-625
6611	01.01.1993	31.12.2006	Feuerkogel (Tawes)	Upper Austria	500-550
6620	01.01.1983	31.12.2006	Gmunden	Upper Austria	625-650
6411	01.01.1986	31.12.1997	Mattsee	Salzburg	625-650
6415	01.01.1999	31.12.2006	Mattsee	Salzburg	625-650
12322	01.01.1985	31.12.2006	Zell am See	Salzburg	550-600

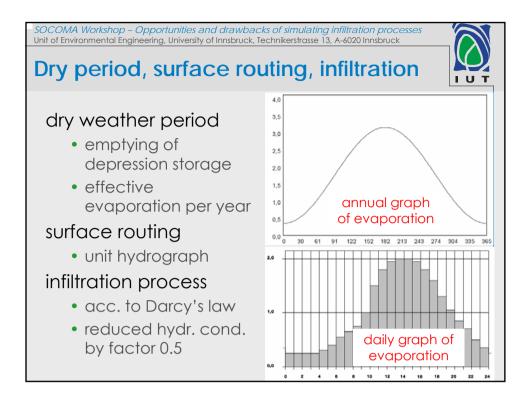
SOCOMA Workshop – Opportunities and drawbacks of simulating infiltration processes Unit of Environmental Engineering, University of Innsbruck, Technikerstrasse 13, A-6020 Innsbruck

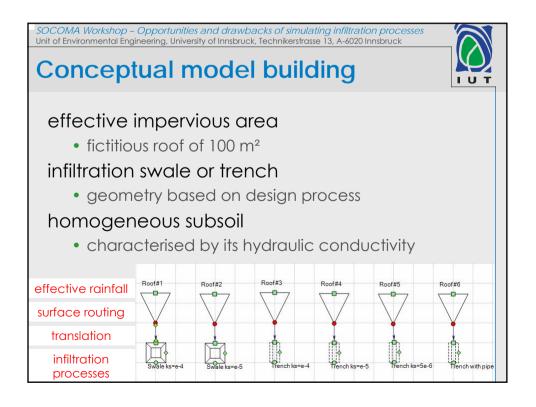
## Pluviographs in detail

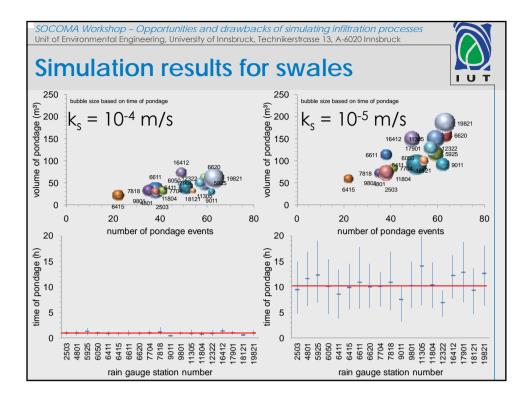


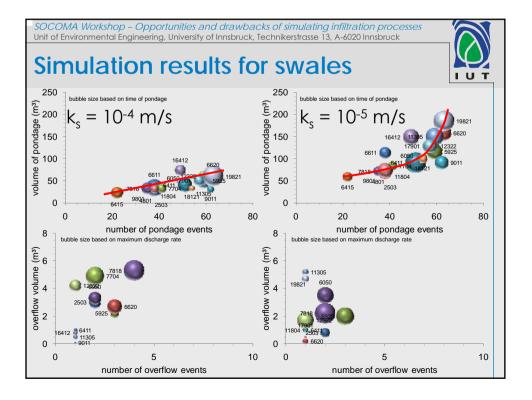
station	begin	end	location	state	evaporation
9801	01.01.1993	31.12.2006	Aigen/Ennstal	Styria	600-625
16412	01.01.1989	31.12.2006	Graz-Universität	Styria	625-650
11804	01.01.1992	31.12.2006	Innsbruck-Univ.	Tyrol	600-625
17901	01.01.1986	31.12.2006	Lienz	Tyrol	600-625
9011	01.01.1984	31.12.2006	Oberndorf/Ebbs	Tyrol	600-625
11305	01.01.1985	31.12.2006	Warth	Vorarlberg	300-400
5925	01.01.1985	31.12.2006	Wien-Innere Stadt	Vienna	625-650

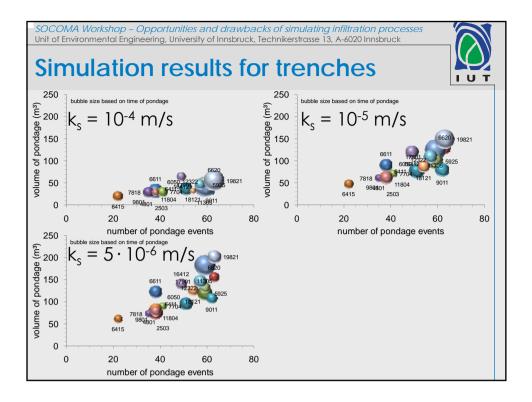


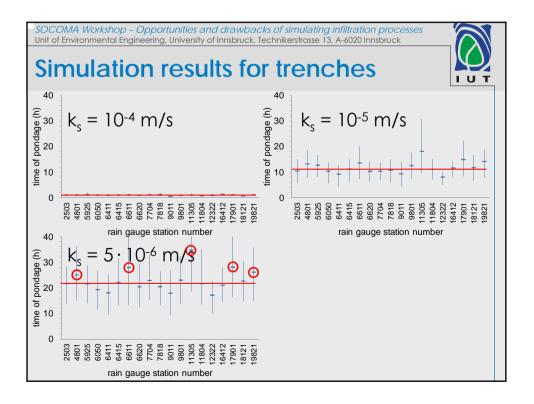


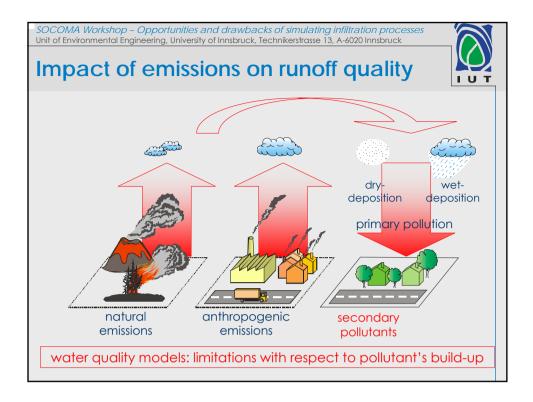










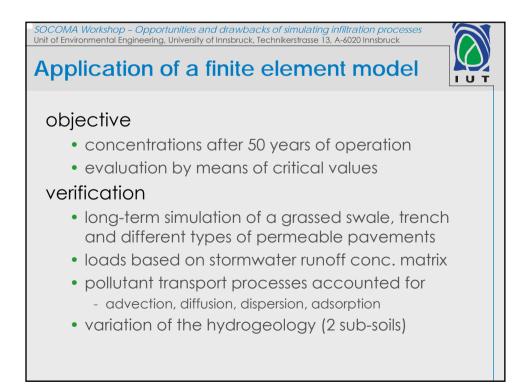


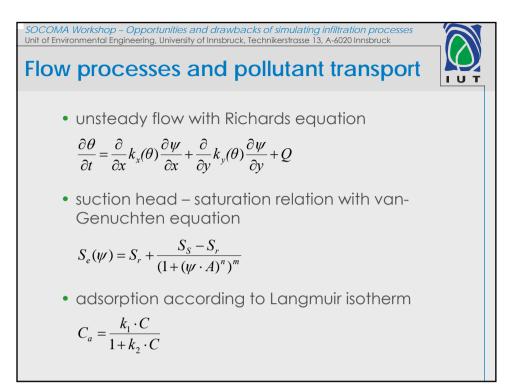
U	nit of En	vironme mv	vate	eering, UI	niversity o	f Innsbruc	nce	entra	13, A-602 atio	<u>n m</u>	atri	x 1	UT
_			unsealed surf.		3	roofs	5	6	, ,	° 1	streets 9	10	11
	parameter	unit	gardens, meadows, cultivated lands	roof runoff, pantiles, concrete tiles, fibr. Cement, glas, bitumen, without zinc	roof runoff, pantiles, concrete tiles, fibr. Cement, glas, bitumen, with zinc	4 planted roofs	copper sheets	zinc sheets	bicycle paths, footpaths, yards	car parks	streets in residential areas	main roads	motorways
1 2	el. cond. pH	[uS/cm] [-]	50 5,0	141 5,7	141 5,7	71 7,5	141 5,7	141 5,7	n.d. 7,4	n.d. 7,4	n.d. 7,4	470 7,4	414 7,4
3 4 5	AFS BSB <sub>5</sub> CSB	[mg/l] [mg/l] [mg/l]	12 2 19	43 12 66	43 12 66	n.d. n.d. n.d.	43 12 66	43 12 66	74 n.d. 70	150 11 70	150 11 70	163 11 105	153 32 107
6 7 8	P ges NH4 NO3	[mg/l] [mg/l] [mg/l]	0,09 0,80 1,54	0,22 3,39 2,78	0,22 3,39 2,78	n.d. 1,30 0,59	0,22 3,39 2,78	0,22 3,39 2,78	n.d. n.d. n.d.	0,18 0,1 2,78	0,18 0,1 2,78	0,29 0,9 5,00	0,20 0,5 2,52
9 10 11 12 13 14	Cd Zn Cu Pb Ni Cr	[_9/1] [_9/1] [_9/1] [_9/1] [_9/1] [_9/1]	0,7 80 11 9 2 3	0,8 370 153 69 4 4	0,8 1.851 153 69 4 4	0,1 468 58 6 3 3	0,8 370 2.600 69 4 4	1,0 6.000 153 69 4 4	0,8 585 23 107 n.d. n.d.	1,2 400 80 137 n.d. n.d.	1,6 400 86 137 14 10	1,9 407 97 170 11 11	3,7 345 65 224 27 13
15 16 17 18 19 20	Na Mg Ca K SO4 CI	[mg/l] [mg/l] [mg/l] [mg/l] [mg/l]	2,14 0,18 7,50 0,56 5,46 2,26	n.d. n.d. 10 n.d. 46,71 7,74	n.d. n.d. 10 n.d. 46,71 7,74	n.d. 7 78 7 n.d. n.d.	n.d. n.d. 10 n.d. 46,71 7,74	n.d. n.d. 10 n.d. 46,71 7,74	n.d. n.d. n.d. n.d. n.d. n.d.	18 n.d. n.d. 4 n.d. n.d.	n.d. n.d. n.d. n.d. n.d. n.d.	108 1 31 2 15 106	194 5 37 5 39 159
21 22	PAH Petrol, HC	[_g/] [mg/l]	0,39	0,44	0,44	n.d. n.d.	0,44	0,44	1,00	3,50 0.16	4,50 0,16	1,65 4,17	2,61 4,76

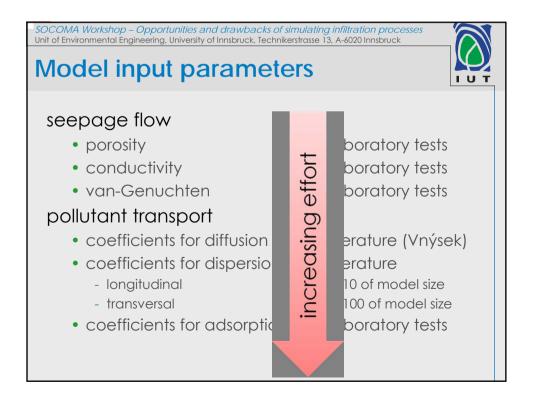
SOCOMA Workshop – Opportunities and drawbacks of simulating infiltration processes Unit of Environmental Engineering, University of Innsbruck, Technikerstrasse 13, A-6020 Innsbruck

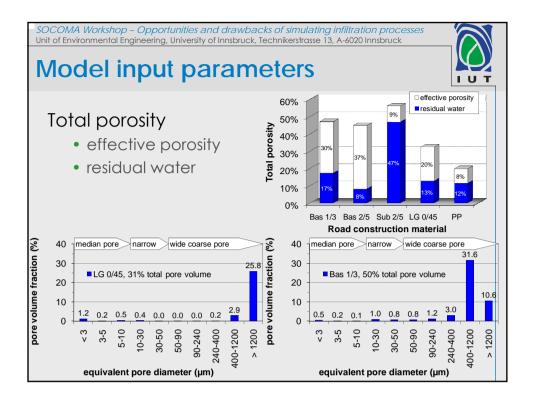
### Constituents in surface runoff as EMC

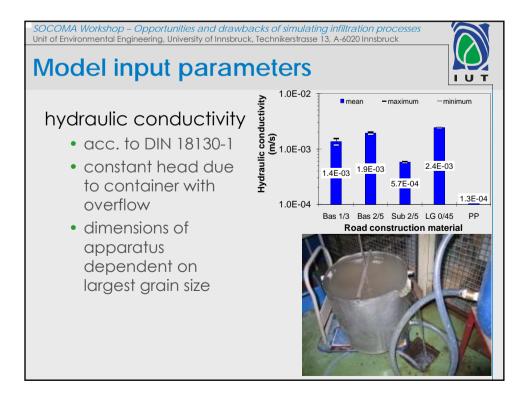
parameter	unit	grassed area	roofs without metals	metal roofs *)	trafficked areas **)
Cd	(µg/L)	0.7	0.8	0.8 – 1.0	0.8 – 1.9
Zn *)	(µg/L)	80	370	370 - 6000	400 - 585
Cu *)	(µg/L)	11	153	153 – 2600	23-97
Pb **)	(µg/L)	9	69	69	107 – 170
PAH <sub>EPA</sub>	(µg/L)	0.39	0.44	0.44	1.0 - 4.5
TSS	(mg/L)	12	43	43	74 – 163
pH-value	(-)	5.0	5.7	5.7	7.4
*) no simultane	eous occurre	ence in roof ru	unoff <sup>**)</sup> origin:	leaded fuel	

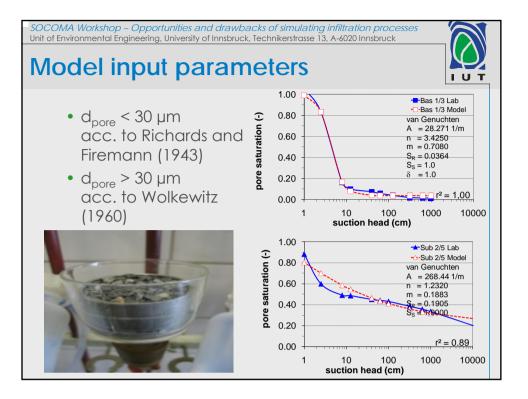


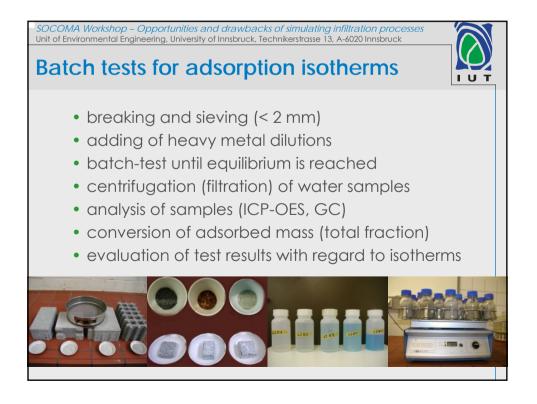


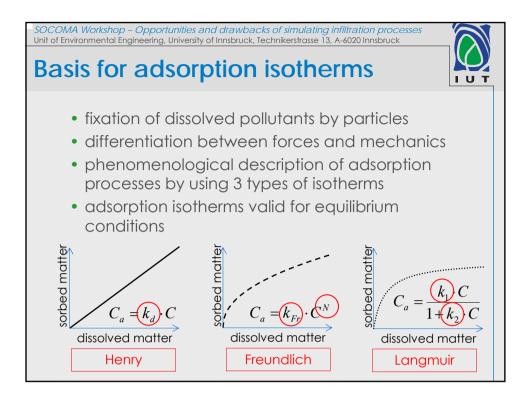


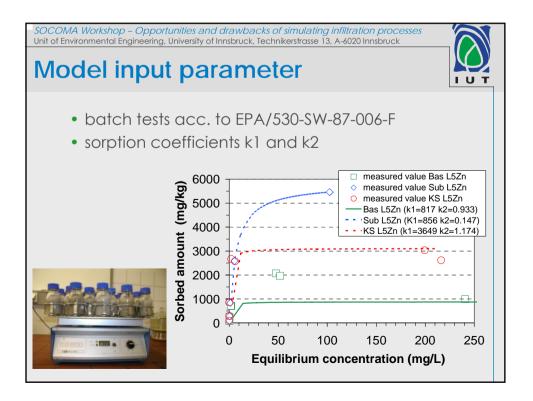


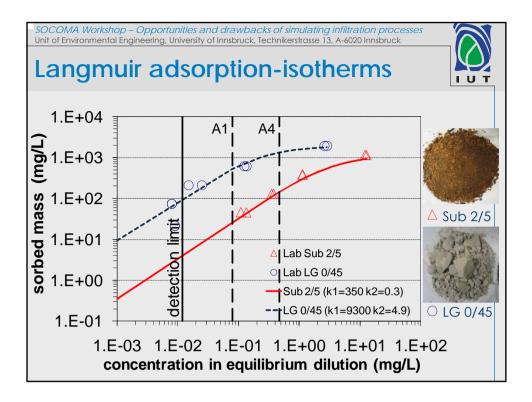


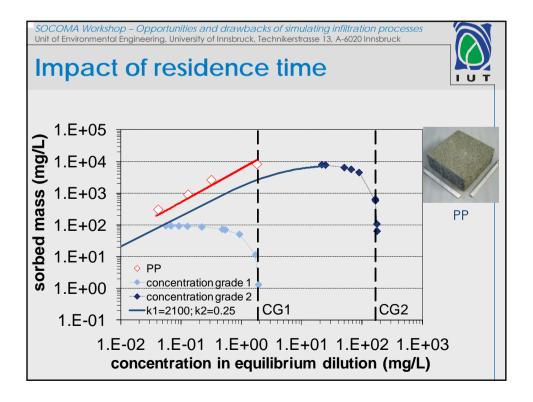


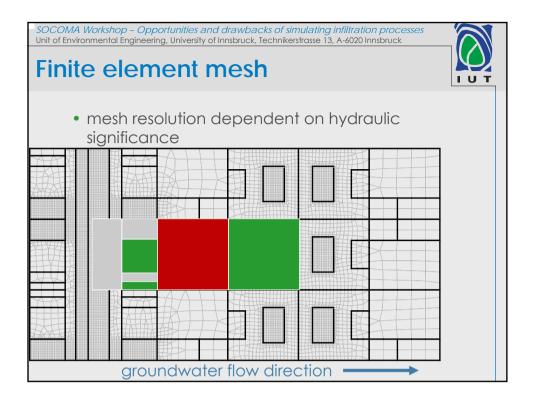


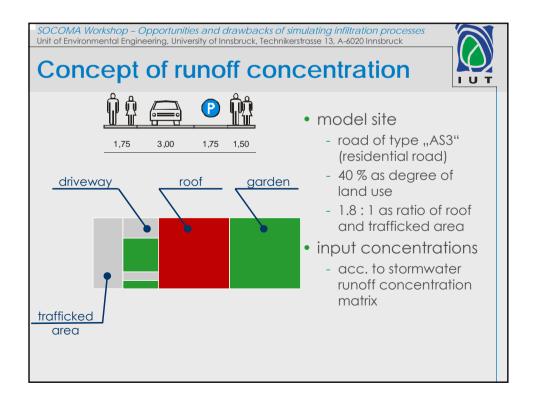


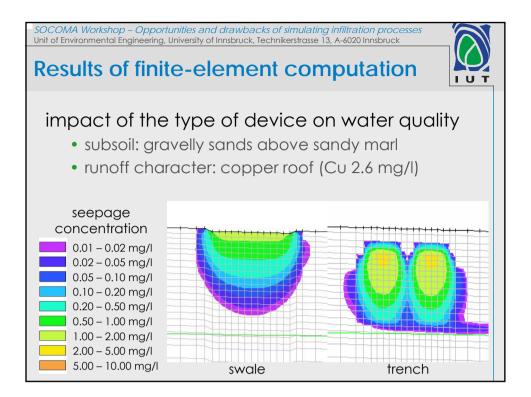


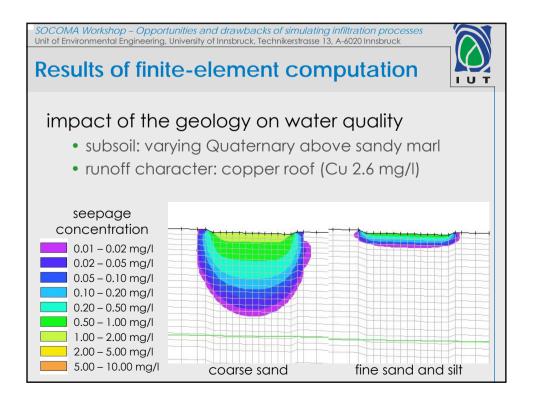


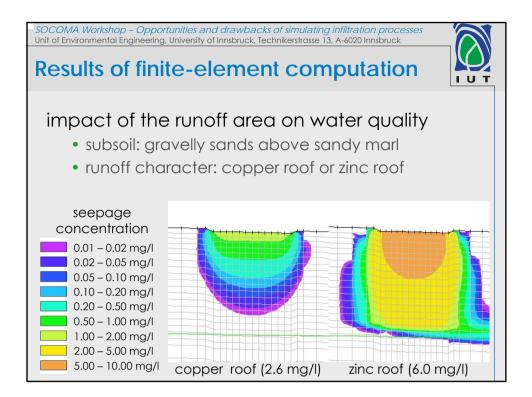




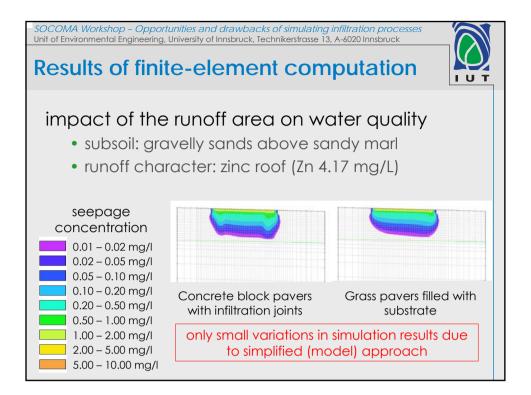


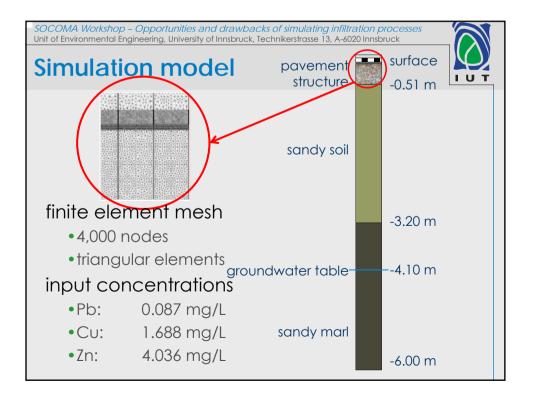






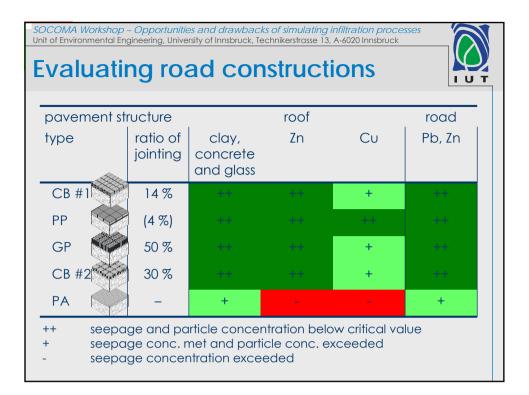
Uni	SOCOMA Workshop – Opportunities and drawbacks of simulating infiltration processes Init of Environmental Engineering, University of Innsbruck, Technikerstrasse 13, A-6020 Innsbruck Simulation results vs. critical values					
2		s vs. cr				
	runoff type		zinc roof	(6.0 mg/l)		
	subsoil characteristics'		fine sand	coarse sand		
	adsorbed mass in first layer of finite mesh	(mg/kg)	78 (60)	99 (60)		
	adsorbed mass 1 m below base of device	(mg/kg)	55 (150)	149 (60)		
	seepage concentration 1 m below device	(µg/l)	190 (500)	5580 (500)		
	seepage concentration acc. to national act	(hð\l)	< 10 (500)	990 (500)		
	depth, where critical value is reached	(m)	~ 2.1	~ 4.0		
	substitution rate for the first 20 cm	(a)	~ 10	~ 2		

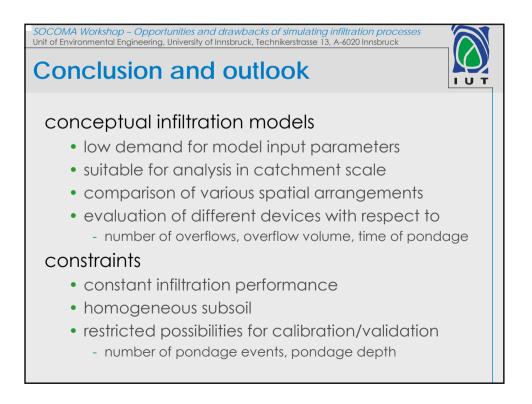


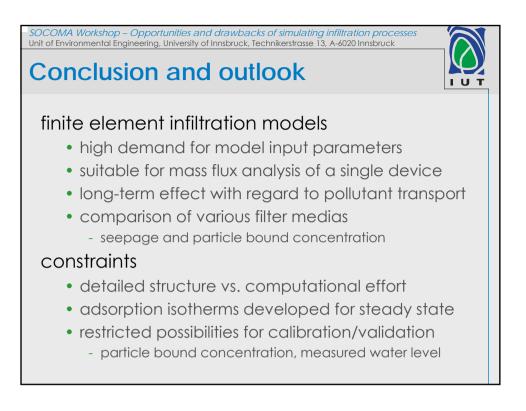


Unit of Environmer	shop – Opportunities , tal Engineering, Universit <b>of pervic</b>	y of Innsbruck, Techniker	strasse 13, A-6020 Innsbru	
60	0 0 0 0 0 0 0 0 0 0 0 0 0 0			
type of paver	blocks with small joints	blocks with large joints	grass paver	porous paver
name	CB #1	CB #2	GP	PP
jointing	basalt 1/3	substrate 2/5	substrate 2/5	basalt 1/3
bedding	basalt 2/5	substrate 2/5	substrate 2/5	basalt 2/5
road base	limestone 0/45	limestone 0/45	limestone 0/45	limestone 0/45

Simulation results for copper						
type of paver	seepage co	oncentration	adsorped p	ollutants		
	below subgrade	unsaturated – saturated soil	in pavement structure	in soil		
	(mg/L)	(mg/L)	(mg/kg)	(mg/kg)		
CB #1	0.288	0.002	135	36		
PP	0.373	0.004	105	51		
GP	0.296	0.003	117	46		
CB #2	0.290	0.002	132	40		









**General literature overview** 

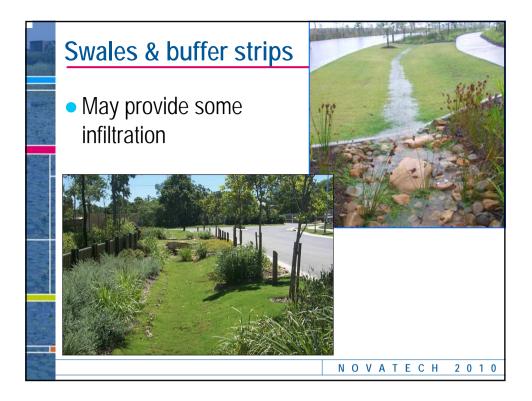
# SOURCE CONTROL PERFORMANCE

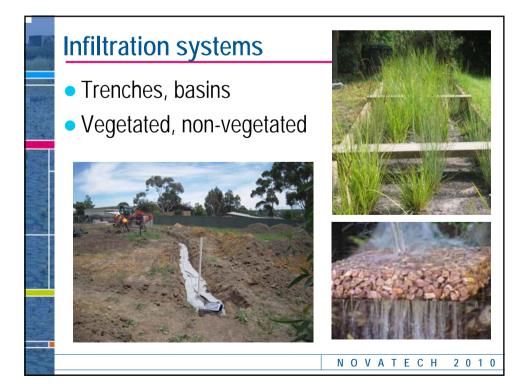
T. Fletcher & P. Hamel

Monash University, Australia



A c	ontinuun	n of techni	ques		Parente Parente Cardinal Cardi
	Detention	Filtration	Infiltration		Retention
Degree of Treatment	Sand filters, mo	Swales, buffer strips lands edia-filtration systems, p nfiltration basins, infiltrati Rain-gardens, biofiltr	on trenches		Rainwater /stormwater harvesting Vegetated roofs
			NOV	ATE	CH 201





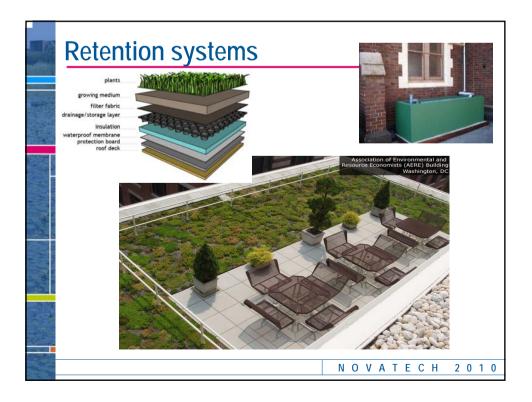
## Filtration systems



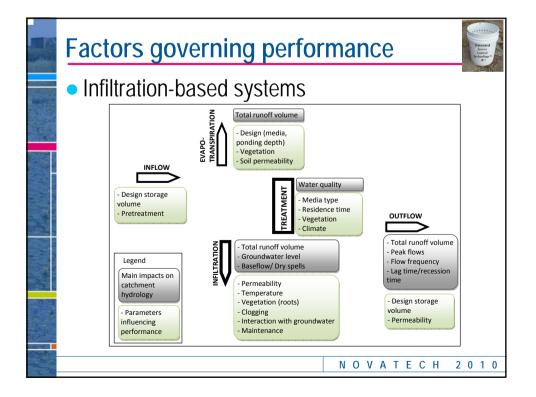
- Vegetated or unvegetated
- Sand filters, media filtration systems
- Discharge stormwater network

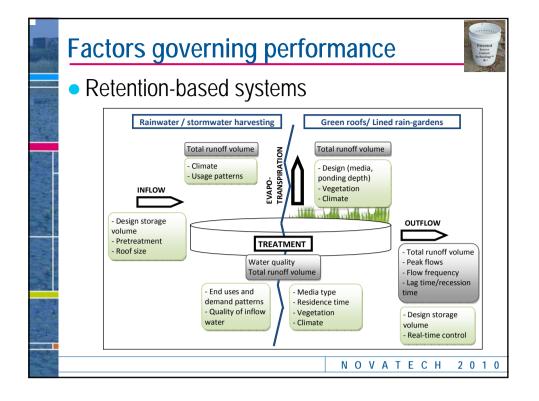


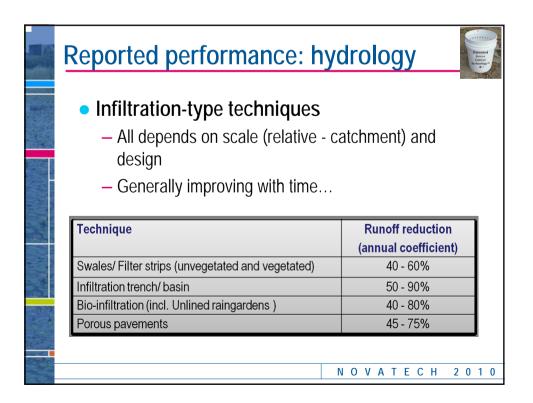


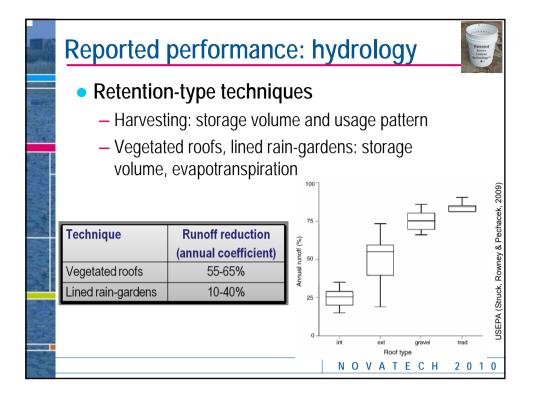


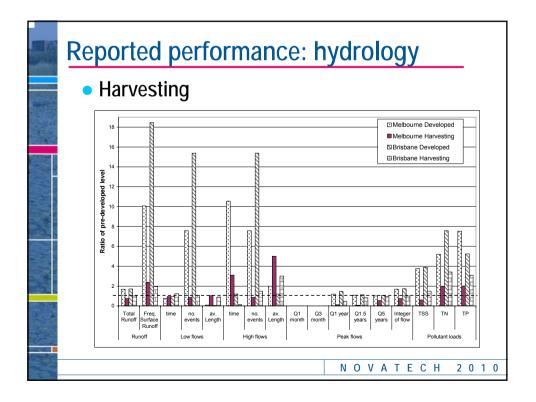












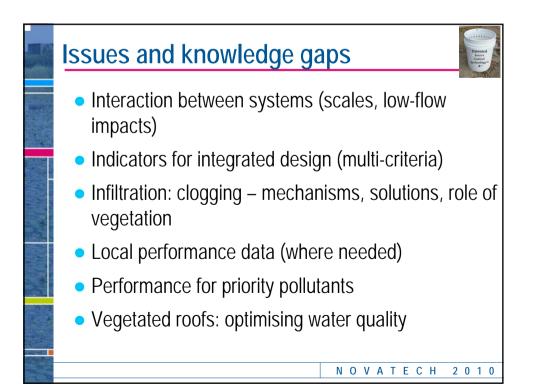
## Reported performance: water quality



2 0 1 0

	Measure	Total N (%)	Total P(%)	TSS (%)	Heavy metals (%)	
Γ	Swales/ Filter strips	25 - 40% <sup>1</sup>	30 - 50% <sup>1</sup>	co. 00%	20 60%	
	(unveg.+veg.)	(55 - 74%) <sup>2</sup>	(52 - 76%) <sup>2</sup>	60 - 80%	20 - 60%	
	Infiltration trench/	50 - 70% <sup>1</sup>	40 - 80% <sup>1</sup>	CE 00%	50 - 95%	
	basin	(57 - 92%) <sup>1</sup>	(63 - 93%) <sup>2</sup>	65 - 99%		
	Biofiltration / bio-	50 - 70% <sup>1</sup>	40 - 80%1	65 000/	50.05%	
	infiltration	(64 - 92%) <sup>2</sup>	(55 - 90%)2	65 - 99%	50 - 95%	
	Porous pavements	60 - 80% <sup>1</sup>	40 - 80% <sup>1</sup>	70 00%	40.000/	
l		(59 - 81%) <sup>2</sup>	(59 - 81%) <sup>2</sup>	70 - 99%	40 - 90%	

NOVATECH



# "GREEN" TECHNOLOGIES & INFRASTRUCTURES FOR THE CONTROL AND TREATMENT OF IMPERVIOUS SURFACE RUNOFF

**Brian Ellis** 

Urban Pollution Research Centre, Middlesex University, UK

# Performance

# The application of vegetative-based BMP source approaches for sustainable urban stormwater watershed planning

### J Bryan ELLIS

Urban Pollution Research Centre, Middlesex University, The Burroughs, Hendon. NW4 4BT. UK. (E-mail: <u>B.Ellis@mdx.ac.uk</u>)

### **1. Introduction**

First-generation surface water drainage followed the conventional practice established by "allto-the-sewer" wastewater conveyance (Chocat et al., 2004) with rainfall-runoff from impermeable urban surfaces being separately sewered. This paradigm argued for the rapid collection and conveyance of impermeable surface runoff with pipes sized according to the Rational formula. Hydraulic and conveyance capacity comprised the driving design criteria and as stormwater flows were considered to be "unpolluted", direct untreated discharges to receiving waters from surface water outfalls (SWOs) were deemed technically appropriate and environmentally acceptable. The increasing awareness of the pollution potential associated with such diffuse, non-point urban discharges led to the introduction of a secondgeneration infrastructure based on best management practices (BMPs) or "sustainable" These alternative techniques have been superimposed onto the drainage systems (SUDS). conventional below-ground sewer drainage system to provide a hybrid solution to address both flow and quality surface water issues. This second-generation drainage approach has been adopted at varying scales and intensities within urban areas across the world, although many alternative BMP drainage technologies have met considerable resistance to their largescale implementation on performance, institutional, legislative and planning grounds (GAO, 2007; Ellis, 2009a). Nevertheless, such BMP approaches currently form the core mitigation philosophy of the US EPA National Pollutant Discharge Elimination System (NPDES) and associated permit program under Section 402 and 101 requirements of the Clean Water Act (CWA) as well as being proposed as essential drainage elements in UK development planning for pluvial flooding and pollution control (CLG, 2006).

However the approach remains essentially based on piecemeal site development and primarily driven by hydraulic requirements of peak storm volume attenuation, storage and treatment, rather than by integrated, catchment-based ecosystem precepts. The US EPA are now reviewing their NPDES policy direction and are considering the adoption of a watershed-based approach for NPDES permits (EPA, 2007) along the lines already contained within the European Union Water Framework Directive (WFD). This shift in philosophy is consonant with a move towards "green infrastructure" thinking in urban stormwater management (Novotny and Brown, 2007). The provision of a "green water" resource base would provide a tangible link with ecosystem services of direct value to the urban community but requires

appropriate land use planning controls with strategic spatial planning approaches applied to both urban development and regulatory policy. Such approaches constitute a major rethinking and re-orientation for established second generation BMP drainage philosophy such that a third-generation of surface water drainage systems may be needed to fully and satisfactorily address this need for a more sustainable, holistic watershed-scale solution (Ellis, 2009b; Marsalek and Schreier, 2009). This paper considers the need for, and basis of, such a new management and planning direction and illustrates some of the advantages and opportunities that might accrue from mitigating approaches that more closely mimic the preurban water balance. Such approaches require a better understanding of hydrologic processes and urban land use factors in order to achieve more sustainable watershed planning.

#### 1. The impermeable surface model

The traditional growth model for urban expansion depicts an exponential relationship between the expansion of the impervious surface and receiving water runoff volume and quality (Figure 1). There is also a widespread belief that the onset of impacted conditions as measured by biotic diversity commences at a baseline of between 10% - 15% impervious cover (IC) with non-sustaining, in-stream aquatic ecologies being evidenced at impervious levels exceeding 40% - 45% (Brabec et al., 2002). Inspection of Figure 1 would suggest that even a threshold of 25% IC might lead to severely impacted conditions under which receiving water quality and aquatic biodiversity becomes significantly depressed. The traditional impervious model, particularly when based on effective impervious area (EIA), also assumes that permeable areas such as open spaces, parkland, gardens etc., do not contribute to runoff. However, such impervious cover modeling predictions are highly generalised and cannot be universally applied to all urban receiving waters (CWP, 2003). Prior water diversions, riparian alterations and land drainage works are all likely to already have degraded the "greenfield" site before development whilst low gradients (<1%) ameliorate the negative effects of initial impermeable cover. On the other hand, there is evidence that riparian woodland, shrub and other vegetative growth tends to mitigate the impact of impervious cover and this appears to be particularly the case for geomorphic and biodiversity indicators which do not degrade much below 15% IC (Cianfranci et al., 2006) in the presence of such vegetation cover.

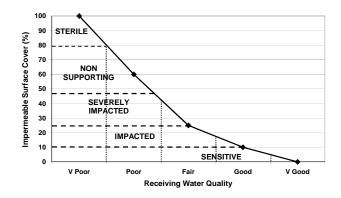


Figure 1. The impervious cover model

There is considerable supporting evidence that a significant riparian vegetative cover serves at least to suppress the onset effect of urban growth on receiving waters by up to some 15% - 20% impervious cover (Wang *et al.*, 2001). Aquatic ecosystems appear to be much more sensitively impacted by the yield of and exposure to toxic sediment within the contributing sub-watershed(s) than by most other urban growth factors. Watershed metrics other than impervious cover, such as percentage vegetative cover and open space or road density might prove to be more appropriate indicators of urbanization, with vegetated cover being effectively the reciprocal of impervious cover. The outcomes of second-generation BMP systems on the receiving water flow, quality and ecological regime need to be considerable to be detected given the statistical variability in the impervious cover model, especially at IC levels less than 10% - 20%. However, pollutant load reductions by BMPs in sub-catchment receiving waters can be normally detected as long as the IC does not exceed 30% - 40% (CWP, 2003).

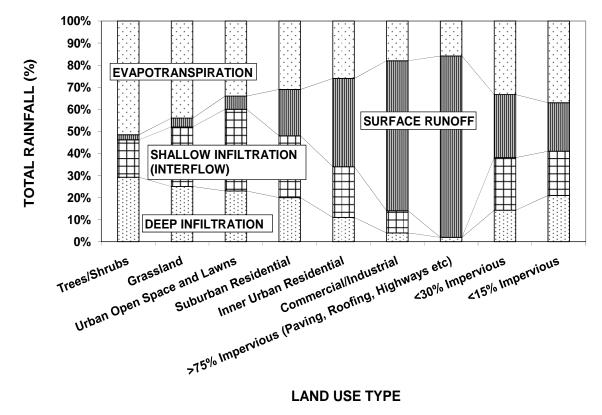
The confirmatory evidence for receiving water improvements following BMP introduction for the control of impervious surface runoff is both scarce and contentious. Washington state for example, has introduced over many years, like many other US states, a widespread BMP approach for the mitigation of stormwater impacts, but most second-generation structural devices have proved inadequate to prevent downstream channel erosion, despite increasingly restrictive designs (Bath et al., 2002). It is argued that structural retrofits to urban drainage have also been largely unable to restore pre-development flows or habitat regimes, and that the fundamental cause of aquatic degradation is the conversion, even at very low levels of impermeable surface cover, of riparian forest, woodland and grassland. Woodland and shrub vegetation together with the introduction of riparian buffer zones combined with carefully optimized design of both storage (to control peak flows) and infiltration (for recharge) BMP facilities are seen as the only means of providing aquatic ecosystem protection in the presence of increasing impermeability (Horner *et al.*, 2001; CWP, 2003; Cappiella *et al.*, 2005).

### 2. The urban water balance

### 2.1 Effect of the impervious surface

The extension of impermeable surface cover produces an impact on the local water balance with surface sealing increasing the active effective-runoff area, whilst at the same time increasing surface compaction also decreases the infiltrative capacity. This process relationship is traditionally expressed in the urban water balance equation as: Precipitation (Rf) = Evapotranspiration (ET) + Infiltration (INF) + Impermeable Surface Runoff (IMPr),where IMPr normally refers to the effective impervious area (EIA) of the watershed. Compaction to soil bulk densities exceeding 1.5 gm/cm<sup>3</sup> can be considered as an inevitable side effect of urbanization with surface "sealing" resulting from a combination of original development compaction, vehicle parking and pedestrianisation (USDA, 2001). Urban open spaces, highway verges, parks and playing fields are likely to have higher impermeabilities than "natural" open spaces in rural surroundings, and this can result in substantial contributions from wetted "pervious" areas to overland exceedance flows during extreme storm events. Under these exceedance conditions, the IMPr term in the urban waterbalance equation exceeds the total impervious area (TIA). Such "pervious" flow volumes and overland routing have yet to be fully considered within the majority of urban runoff models. Surface sealing and compaction will therefore result in increased overland runoff flow with runoff volume being generally linearly related to urban land use type and activity as illustrated in Figure 2 based on water balance studies in the UK and Germany (Ellis, 2009b) and confirmed by very similar results reported for Canadian cities (Marsalek and Schreir, 2009).

It has been argued that such hydrological adjustments also result in shorter surface detention times, thus diminishing both evaporation and groundwater recharge. It is certainly the case that reductions in interflow, shallow and deep infiltration processes have characteristic reducing footprints in the urban water balance as impermeable cover increases (EPA, 2005), and this is evident from inspection of Figure 2. However, as also seen from the figure, both shallow and deep infiltration as well as evapotranspiration (ET) remain substantial components of the overall water balance at impervious covers less than 35% or so. This is largely due to the significant depression storage and initial losses that can occur on impermeable urban surfaces and which have been underestimated in most urban runoff modeling studies (Brabec *et al.*, 2002; WaPUG, 2004).

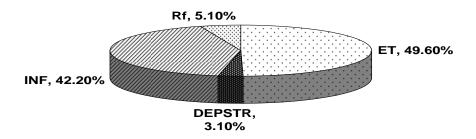


### Figure 2. The urban water balance.

Macro (surface roughness), meso (surface puddles) and macro-storage (total connectivity to the drain) components of overall depression storage can add substantially to initial losses (Ellis and Revitt, 2008). Studies in metropolitan Manchester and London in the UK have indicated that residential land use covers up to 50% of the total urban area with medium density housing accounting for some 37% (Gill *et al.*, 2008). It is within this latter dominant category that 32% of potential ET surfaces are found e.g lawns, hedges, grass verges, shrubs, trees etc; features which are not commonly represented by traditionally mapping approaches. Lack of inclusion of these types of surface cover will underestimate the ET and INF losses from total impermeable surface runoff. The Grimmond and Oke (1986) model also predicts that "isolated" urban depression storage patches will lose water at enhanced evaporative rates due to the provision of sensible heat from surrounding drier (and frequently hotter) impermeable surfaces. In addition, strong mechanical wind turbulence set up by the uneven

urban building profile and boundary conditions results in a vigorous surface-boundary layer effect increasing the evaporation potential.

There is also long standing field evidence that so-called impermeable surfaces can infiltrate significant quantities of both surface water and micro-pollutants (Ellis and Harrop, 1984; Hollis, 1997). This is particularly the case for bitumen (blacktop) surfaces where pore openings, potholes, hollows and cracks develop, especially along and adjacent to roadside gutter channels. Irrigation experiments have indicated that up to 80% of applied surface water on asphaltic, low trafficked street surfaces does not find its way to the gullypot but is either held on the surface or infiltrates into the sub-grade (Hollis and Ovenden, 1988); peak runoff showed an attenuation of 24% for 5 minute rainfall intensities. Laboratory testing of differing paving materials has confirmed the high evaporative and infiltrative capabilities of bitumen surfaces, primarily resulting from depression storage and subsequent evapotranspiration and infiltration (Mansell and Rollet, 2007). Figure 3 shows the distribution of total monthly rainfall from isolated small rainfall events.



# Figure 3. Percentage distribution of water balance components for storm events of <1:5 return interval (RI).

Thus apparently impermeable surfaces can be subject to high and variable losses and therefore conventional water balance modeling of rainfall-runoff overestimates for frequent, low magnitude storm events (<1:5 RI) but also significantly underestimates runoff volumes for infrequent, high magnitude extreme events (>1:30 RI). This has implications for BMP planning which frequently results in over-sizing of treatment devices and particularly in the case of storage BMPs such as detention/retention ponds and wetlands.

### 2.2 The influence of vegetative surfaces

Streets and highways comprise the most significant flow and pollutant sources and conduits within the urban environment, and serve as primary conveyance pathways to move surface water from rooftops, lawns, driveways, pavements/sidewalks and the street surface itself. Street edge alternative (SEA) retrofitting involving re-vegetation of the impermeable urban surface, not only can attenuate and phyto-remedially treat the final cumulative impermeable surface runoff at source, but can also improve neighbourhood aesthetics, calm traffic flows and act as focal educational centerpieces. Roadside SEA retrofits are ideal planning alternatives in that they deal with the most polluted fractions of impermeable surface runoff whilst at the same time local costs are minimized since the retrofit is located in the dedicated right-of-way. In addition, vegetative introductions may be a viable option for runoff pre-treatment if structural retrofits prove not to be feasible. Biofiltration alternatives such as rain

gardens, street planters, tree pits, pocket wetlands, buffer (filter) strips and modified swale channels will enhance both evapotranspiration and infiltration rates such that a100% retention of small rainfall events (<1:1 RI) can be achieved. Field monitoring of SEAs in Seattle, Washington state have observed up to 99% reduction in total volumes of stormwater discharging annually from the street surface (Horner *et al.*, 2002). The capture of such high levels of runoff volume will also give a complementary reduction of up to 65% - 75% of total suspended solids on a long term average basis (Hunt and Lord, 2006).

Such rain gardens, street planters, tree pits and biofiltration features have become common design elements in Low Impact Design (LID) approaches and can make substantial contributions in re-aligning the urban water balance towards "natural" pre-development conditions (Ellis et al., 2004). As yet there have been few studies to attribute the observed rainfall losses to either of evaporative, interflow, shallow or deep groundwater receptors. However, there is an implicit assumption in the prevailing studies of these bioretention facilities that the majority pathway is by infiltration to groundwater (Dietz and Clausen, 2005), hence the recommendation for lining and/or underdrains where there is a potential threat from toxic first-flush pollutants from the road surface and for a minimum 250 - 300 mm amended topsoil cover (Pitt *et al.*, 2005). The previous discussion of the urban water balance distribution however, suggests that this assumption may be incorrect, as the dominant losses appear to be associated with ET following on from initial depression storage.

Small scale bioretention retrofits can also be implemented for cycle tracks, play and recreational areas, driveways and playing fields. These retrofits may comprise grass filter (buffer) strips, grass channels or small infiltration and porous paving devices, and may only make up some 5% or so of the total impervious area within the urban sub-catchment. However, they can be readily retrofitted, are low cost options and can help to solve site drainage problems. Vehicle parking areas (> 1.5 - 2.0 ha) also offer good biofiltration opportunities (e.g tree pits, planters, mixed woodland/shrubs, grass filter strips, vegetated swales), especially as more recently constructed parking sites have more generous setbacks for screening, landscaping, noise reduction etc... Perimeter and island bioretention approaches, together with porous surfacing, can also be used very effectively on smaller vehicle parking sites. Within the larger urban sub-catchment, there is considerable evidence now available to demonstrate the significant reduction in runoff that can be achieved by the introduction of site vegetation and riparian corridors, particularly for low intensity, short duration rainfall events, as evidenced for example in the Baltimore studies of Wang et al (2008) or the Vancouver investigations of Asadian and Weiler (2009). The latter studies indicated an urban canopy interception varying between 50% to 61%, corresponding to a net annual loss of 20 – 32 mm of rainfall-runoff.

It is also clear that ET may well be the major factor in determining how much water is available for infiltration and as such, represents the controlling component of the urban water balance profile. Effective stormwater BMP design needs to consider the relative significance of the various water balance processes as well as their seasonal variations. Design approaches which aim for landscape or watershed-scale infiltration which preserve the natural ET pump mechanism as much as possible would considerably improve overall performance. This objective would be supported by the introduction of two- (or even three-) tier vegetation for biofiltration BMPs to maximize the canopy cover. The adoption of a dense, multi-layered vegetation (grass, shrubs and trees) for these BMPs would represent a fundamental shift in the existing rather cosmetic design guidelines for second generation drainage systems.

## 3. A third generation drainage approach?

### 3.1 Basis and need for a new approach

Given the contentions and speculations associated with the traditional impervious area cover model and the analysis of the urban water balance, what then are the "best" management practices to provide receiving stream and channel protection as well as water quality and recharge benefits? The foregoing analysis would suggest that there are critical urban land use components which are necessary if a more natural and sustainable water balance and drainage system is to be achieved:

- minimization of impervious areas, local land disturbance and surface compaction. This planning precept is to some extent being recognized with for example, new stormwater regulations for Maryland, US requiring that all redevelopment sites reduce existing impervious surfaces by at least 50%.
- (re)-introduction and preservation of site and sub-catchment vegetation cover and open space,
- preservation of critical ecological zones such as riparian corridors, wooded wetlands, floodplains etc..

These three primary land use strategies offer basic planning steps to the restoration of an effective urban water balance, the restoral of watershed functions and a framework to ensure adequate water resource protection (Cappiella *et al.*, 2008).

The majority of national BMP guidance provide broadly similar design outcomes despite major advances in BMP research that has taken place over the past 5 to 10 years. The majority of design guidance manuals derive narrative tables for the water quantity and quality effectiveness of differing stormwater management practices for different pollutant groups and water balance re-adjustments e.g groundwater recharge, peak runoff and volume controls. Other potential secondary benefits e.g landscape enhancement, recreation/amenity, community acceptance, safety etc., are also often superimposed in the decision-path analysis to identify a suite of recommended BMP devices. The similarity of design guidance outcomes and recommendations can be interpreted as implying that many BMPs being installed may not be applying the most effective or innovative technology and that far too many poorly or under-performing BMPs exist as well as being overdesigned in many cases in terms of water quality performance. This conclusion is confirmed to some extent by reports of failure or under-performing BMP controls, particularly infiltration systems (Schluter and Jefferies, 2005).

A new design approach paradigm may be needed which considers the effect of impervious cover growth from rooftop to the receiving water with the approach specifically aimed at the restoration of more "natural" pre-development water balance conditions (Schueler, 2004). This third generation approach should focus on better low impact site design (LID or cluster development), with a substantial preservation and/or re-introduction of vegetative cover (with a target minimum 30% - 40% canopy cover) within the development, infill or retrofit (Ellis *et al.*, 2004). Such percentage canopy cover can increase interception by up to 40% of the total rainfall event depth, retaining between  $0.5 - 3 \text{ m}^3$ /day and maintaining infiltration rates of up to 30 cm/hour. This level of canopy cover is important to maximize interception, ET and pollutant phytoremediation rates as well as providing shade for the impermeable surface, thus ameliorating the "heat island" effect. The basic aim should be to restore or maintain the original pre-development water balance as much as possible such that vegetative cover should be capable of providing at least the first 10 - 15 mm of stormwater retention and treatment. This would then enable second generation BMPs to offer a more effective attenuation and

"polishing" function over a wider range of storm event return periods, providing greater capability for enhanced and sustained receiving water health.

## 3.2 Vegetation-based source BMPs

As previously argued, streets and associated impermeable surfaces comprise the major source of runoff volume and pollutant loads and it is this general land use category on which the planning focus of BMP approaches needs to be placed to achieve a fully sustainable urban drainage system. One significant vegetative control device incorporated into LID practice is the rain garden or street planter and which is sometimes termed a pocket wetland in European practice. Rain gardens represent a scaled-down combination of infiltration and biofiltration devices. Stormwater runoff from the impermeable surface is diverted into a local hollow where it can percolate through an organic filter medium such as a compost or amended topsoil layer (Pitt *et al.*, 2005). An overflow spillway or overland flow path normally allows for larger storm events which exceed the filtration and retention capacity, although overflow can be by means of a surface pipe. Some of the collected stormwater will percolate down to contribute to interflow and shallow groundwater, with the remainder being discharged off-site via underdrains. In vulnerable groundwater zones, an impermeable geotextile liner can be introduced to prevent infiltration into the underlying unsaturated zone.

A dense low vegetation cover, low water velocities and extended retention times to enhance evapotranspiration are required to ensure performance effectiveness. Curb-cuts can allow exceedance water to flow "downstream" from one rain garden to the next, or to facilitate discharge into an adjacent swale or back into the road gutter channel for entry into a conventional roadside inlet. Such consecutive rain garden cells essentially comprise a cellular treatment train and can achieve high levels of vegetative "canopy" cover within the impermeable area. Such elongated vegetated cells have a long flow path length and high surface roughness along the flow path which increases the time of concentration and residence times. The 2ndAvNW Seattle SEA fully contain and convey the 1:25, 24 hour duration storm event and are capable of detaining up to 60 m<sup>3</sup> within the conveyance system. The SEA is estimated to reduce the IC for the 1 ha residential block by at least 10% (Horner *et al.*, 2002).

The utilisation of curb extension planting into SEA design, as well as pavement/sidewalk tree pits and trenches can be readily incorporated into the green street approach with the extensions also providing traffic control measures through reductions in road width. Tree pits serve as mini-detention "puddles" capturing runoff from the paved surface. The use of tree pits set in permeable pavements with a continuous soil trench connecting the pits under the sidewalk, provides a shared soil volume for the vegetation and infiltration devices. Disconnected roof leaders/downspouts can also discharge to such pavement pits and trenches. However, such bioretention facilities need to be used conjunctively in an integrated planning design as they individually have a limited hydraulic capacity. High density cluster development and vegetative landscaping needs to be accompanied by extensive rooftop disconnection, green roofs and where feasible and appropriate, rainwater harvesting (tank storage).

### 3.3 Costs and performance

Construction costs per hectare of build type need to be identified which incorporate such plot-based biofiltration BMPs and which should comprise an integral element of normal planning design for urban surface drainage. One US study suggests that rain garden retrofits cost (at 2006 prices), between US140 - 1000 per m<sup>3</sup> stormwater treated, tree pits  $250/m^3$  and grass swales/filter strips  $400 - 550/m^3$  compared to  $400 - 1300/m^3$  for green roofing

(Schueler *et al*, 2007). Grass channel (swale) performance can be as high as 70% - 80% when combined with a high (>70%) vegetative cover and will also require less efficient underdrainage (Hunt and Lord, 2006).

Such bioretention systems can reduce runoff volume by anything between 25% - 60% as a result of evapotranspiration (ET) and infiltration (INF) loss. The key factors in such reductions are soil type and depth as well as local hydraulic gradients. However, adequate and regular vegetation management would also be a key maintenance task for sustained system performance. Post-development instability within the contributing drainage area, poor soil media and adverse elevation and gradients can all affect on-going performance. In areas of potential groundwater vulnerability, infiltrative BMP systems should require mandatory underdrain and surface overflow facilities. The inclusion of green roofs and downspout disconnection into urban green infrastructure would also provide average site volume reductions of up to 34% with peak volume reductions of up to 60% for storm events up to the 1:10 RI period. The combination of such plot and site-based management practices could reduce sediment and bacteria yields by up to 90%, metals by up to 65% and nutrients to between 35% - 60% as well as offering substantial runoff control and help restore urban receiving water ecology (Walsh et al., 2005). However, it must be remembered that street runoff can discharge elevated levels of micropollutants such as metals and hydrocarbons which can accumulate on vegetation surfaces and in the surface soil layers of SEA BMPs. Theses biofiltration devices should be maintained in dense vegetation to prevent dust generation and surface soils may need landfill disposal. H

Schueler *et al* (2007) contend that such site-based, integrated micro-biofiltration approaches can address and protect flood, water quality and in-stream erosion objectives up to as much as 40% impermeable cover levels and that they might even be feasible up to 60% IC levels depending on local circumstances. However, to ensure long term performance effectiveness, these vegetative controls should be complemented by optimization of existing site and end-of-pipe second generation BMPs, together with the implementation of advanced treatment techniques to deal with pollution "hotspots", all of which require strict regulatory and planning controls to be successful. Community surveys of resident demand for, and use of, open/green spaces and green infrastructure would also help support municipal planning programs.

#### 3.4 "Leaf-Out" inventory and analysis

Inventories of existing as well as potential vegetative and open space areas within an urban sub-catchment can be derived from GIS survey and/or satellite imagery. Cappiella *et al.*, (2005) have described this quantification of potential vegetative canopy cover as "leaf-out" analysis. Such areas would include gardens, parks, playing fields, institutional grounds, bare, derelict and vacant ground, as well as other open space in addition to existing wooded and shrub areas, all of which constitute potential re-vegetation locations. Future planned land use zoning and site development can be superimposed on the leaf-out analysis to identify locations that may be candidates for the introduction of biofiltration SEAs, green roofs, small scale urban forestry, green corridors etc. Adjustments to runoff coefficients resulting from "leaf-out" third generation BMP implementation can be estimated from upgraded water balance models to determine their potential effects upon total and peak runoff volumes as well as pollutant loadings and to assess their effects upon the local water balance (Viavattene *et al.*, 2008). The results would also help support the predictions of receiving stream health and ecosystem survival for watershed planning and to support local community environmental stewardship.

#### 4. Conclusions

A feasible strategy to counteract the negative effects of increasing impermeable cover is to develop roof and street vegetative BMPs together with green riparian corridors within subcatchments, linking open space fragments to increase biodiversity and migratory capacity of both flora and fauna. Such corridors will also provide a connection from the site to the watershed scale as well as offering opportunities for the promotion of local community environmental stewardship. The widespread implementation of roof and SEA retrofits and riparian corridor vegetation together with BMP optimization would provide an appropriate environmental infrastructure framework within which such stewardship could be fostered.

The application of a multi-layered, sub-watershed based approach to urban drainage which jointly and concurrently considers both plot and site infrastructure is needed to enhance the implementation and performance of second generation stand-alone and/or treatment train BMPs. A more intensive planning and regulatory approach which emphasises the importance of retaining or re-establishing the original water balance through reduction or control of impervious cover effects using vegetative surface controls can offer a further means of achieving sustainable urban watershed management. Such passive planning manipulation of the urban micro-climate and water balance through extending green space and infrastructure to maximize evapotranspiration and infiltration processes, can derive clear returns for community investment.

#### References

- Asadian, Y and Weiler, M. (2009). A new approach in measuring rainfall interception by urban trees in coastal British Columbia. *Water Qual Res J Canada.* 44 (1), 16 25.
- Bath, D.N., Hartley, D and Jackson, R. (2002). Forest cover, impervious surface area and the mitigation of stormwater impacts. *J Am Water Resources Ass.*, *38* (3), 835 845.
- Brabec, E., Schulte, S and Richards, P. (2002). Impervious surfaces and water quality: A review of current literature and its implications for watershed planning. *J Plan Lit*, *16*, 499 514.
- Cappiella, K., Wright, T and Schueler, T. (2005). Urban Watershed Forestry Manual. Vols 1 3, NA-TF-04-05, US Dept of Agriculture (USDA), Newtown Square, Pennsylvania. US.
- Cappiella, K., Collins, K., Hirschman, D and Novotney, M. (2008). New approaches to "greening" stormwater. *Proc. Water Envir. Federation., Sustainability2008.* 658 674.
- Chocat, B., Ashley, R., Marsalek, J., Matos, M.R and Rauch, W. (2004). Urban drainage: Out-of-sight-out-of-mind? Proc .5<sup>th</sup> Int Conf NOVATECH04. Sustainable Techniques and Strategies in Urban Water Management. (pp. 1659 – 1690). July 2004. GRAIE, Villeurbanne Cedex, Lyon. France. ISBN 2950933769.
- Cianfrani, C. M., Hession, W. C and Rizzo, D. M. (2006). Watershed imperviousness impacts on stream channel condition in SE Pennsylvania. *J Am Water Resources Ass.*, 42 (4), 941–956.
- CLG. (2006). Development and Flood Risk. Planning Policy Statement 25 (PPS25). Communities and Local Government. London. UK.
- CWP. (2003). *Impacts of Impervious Cover on Aquatic Ecosystems*. Watershed Protection Techniques, Monograph 1, Center for Watershed Protection, Ellicott City, Maryland. US.
- Dietz, M.E and Clausen, J.C. (2005). A field evaluation of raingarden flows and pollutant treatment. *Water Air Soil Poll.*, *167* (1/4), 123 138.

- Ellis, J.B. (2009a). Managing urban runoff. In Ferrier, R and Jenkins, A (Edits): *Handbook* of Catchment Management. (Chapter 7, pp. 155 182). Wiley-Blackwell Publishing, London. ISBN 978 1 4051 7122 9.
- Ellis, J.B. (2009b). Third generation urban surface water drainage: From rooftop to the receiving water sub-catchment. *Proc.* 11<sup>th</sup> Int Conf Urban Drainage (ICUD). 31 August – 5 September 2008. Edinburgh, Scotland, UK. CD-ROM. ISBN 978 1899796 212.
- Ellis, J.B and Harrop, O. (1984). Variations in solid loadings to roadside gullypots. *Sci Total Environ.*, *33*, 203 212.
- Ellis, J.B and Revitt, D.M. (2008). Quantifying diffuse pollution sources and loads for environmental quality standards in urban catchments. *Water Air Soil Poll.*, 8 (5/6), 577 585.
- Ellis, J.B., Scholes, L., Revitt, D.M and Oldham, J. (2004). Sustainable approaches for urban development and drainage in the 21<sup>st</sup> century. *J Inst Civil Engineers; Municipal Engineer.* 157 (ME4), 245 250.
- EPA. (2005). National Management Measures Guidance to Control Nonpoint Source Pollution from Urban Areas. EPA 841-B-05-004, US Environment Protection Agency, Office of Water, Washington DC. US.
- EPA. (2007). Watershed-based NPDES Permitting Technical Guidance. Report 833-B-07-004. Office of Wastewater Management. Environment Protection Agency. Washington DC. US.
- AO. (2007). Further Implementation and Better Cost Data Needed to Determine Impact of EPA's Stormwater Program on Communities. Report GAO-07-479. General Accounting Office. Washington DC. US.
- Gill, S.E., Handley, J.F., Ennis, A.R., Pauleit, S., Thenray, N and Lindley, S.J. (2008). Characterising the urban cover of UK cities and towns: A template for landscape planning. *Landscape Urban Plan.* 87 (3), 210 – 222.
- Grimmond, C.S.B and Oke, T.R. (1986) Urban water balance: Results from a suburb of Vancouver, British Columbia. *Water Resources Research.*, 22 (10), 1404 1412.
- Hollis, G.E. (1997). Rain, roads, roofs and runoff: Hydrology in cities. Chapter 15 in Goudie, A. (Edit): *The Human Impact: A Developing Literature*. Blackwell Publishing, London. UK. ISBN 0631199799.
- Hollis, G.E and Ovenden, J.C. (1988). The quantity of stormweater runoff from ten stretches of road, a car park and eight roofs. *Hydrol Process.* 2 (3), 227 243.
- Horner, R.R., May, C., Livingston, E., Blaha, D., Scoggins, M., Tims, J et al. (2001). Structural and non-structural BMPs for protecting streams. In Urbonas, B (Edit): Linking Stormwater BMP Designs and Performance to Receiving Water Impact Mitigation. Proc Engineering Research Foundation Conf., (pp.60 – 77). Snowmass, Colorado. Amer Soc Civil Engineers., Reston, Virginia, US. ISBN 0784406022.
- Horner, R.R., Lim, H and Burges, S.J. (2002). Hydrological Monitoring of the Seattle Ultra-Urban Stormwater Management Projects. Technical Report 170, Water Resources Series, Dept Civil & Environ Engineering., University of Washington, Seattle. US.
- Hunt, W.F and Lord, W.G. (2006). *Bioretention Performance, Design, Construction and Maintenance*. AGW-588-05. Urban Waterways, N Carolina Cooperative Extension Service, State University, Carolina. US.
- Mansell, M and Rollet, F. 2007. The water balanced of paved surfaces in urban areas. *Proc SUDSnet National Conf.* 14 November 2007. University of Coventry, Coventry, UK. (Available at <u>www.sudsnet.abertay.ac.uk)</u>
- Marsalek, J and Schreir, H (Edits). (2009). Innovative Approaches to Stormwater Managment in Canada. *Water Qual Res J Canada*. 44 (1). 110 p.

- Novotny, V and Brown, P (Edits). (2007). *Cities of the Future*. IWA Publishing, London. UK. ISBN 184 3391 368.
- Pitt, R., Cha, S., Clark, S and Lantrip, J. (2005). Soil structure effects associated with urbanisation and the benefits of soil amendments. *Proc World Water & Environ Resources Conf.*, Anchorage, Alaska. Amer Soc Civil Engineers., Reston, Virginia, US.
- Schluter, W and Jefferies, C. (2005). The real issues with in-ground SUDS in Scotland. In: Mikkelson, P.S (Edit): *Proc 10<sup>th</sup> Int Conf on Urban Drainage (ICUD)*. August 2005. Copenhagen, Denmark. IWA Publishing, London. UK. ISBN 978 1843395 744.
- Schueler, T. (2004). An Integrated Framework to Restore Small Urban Watersheds. Small Watershed Restoration Series, Manual 1. Centre for Watershed Protection, Ellicott City, Maryland, US.
- Schueler, T., Hirschmann, D., Novotney, M and Zelinski, J. (2007). *Urban Stormwater Retrofit Practices*. Urban Watershed Restoration Manual No.3., Center for Watershed Protection. Ellicott City, Maryland. US.
- USDA. (2001). Impact of Soil Disturbance During Construction on Bulk Density and Infiltration in Ocean County, New Jersey. National Resources Conservation Service, US Dept of Agriculture, Ocean County Soil Conservation District, Forked River, New Jersey. US.
- Viavattene, C., Scholes, L., Revitt, D.M and Ellis, J.B. (2008). A GIS based decision support system for the implementation of stormwater best management practices. *Proc 11<sup>th</sup> Int Conf on Urban Drainage (ICUD)*. August 2008. Edinburgh, Scotland. CD-ROM. ISBN 978 1899796 212.
- Walsh, C.J., Fletcher, T.D and Ladson, A.R. (2005). Stream restoration in urban catchments through re-designing stormwater systems: Looking to the catchment to save the street. *J N Am Benthol Soc.*, *24* (3), 690 705.
- Wang, L., Lyons, J., Kanehl, P and Bannerman, R. (2001). Impacts of urbanisation on stream habitat and fish across multiple spatial scales. *Environ Manage.*, 28, 255 266.
- Wang, J., Endreny, J.A and Novak, D.J. (2008). Mechanistic simulation of tree effects on the urban water balance model. *J Am Water.Resources Ass.*, 44 (1), 75 85.
- WaPUG. (2004). Urban Rainfall and Runoff. Report of WaPUG Workshop, 30 April 2004, Coleshill, Birmingham. Foundation for Water Research (FWR). Marlow, Bucks. UK.

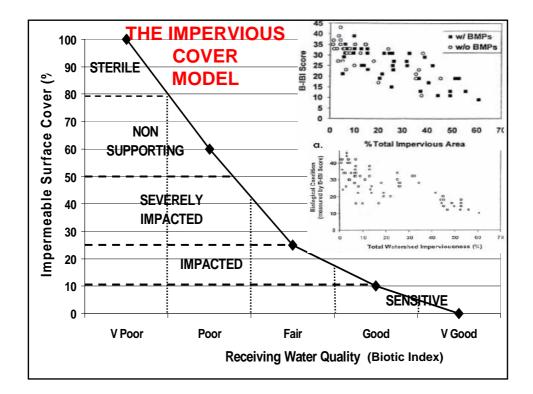
# ELEMENTS IN FAVOUR OF SOURCE CONTROL: THE EXPERIENCE OF THE FRENCH ON-SITE OBSERVATORY OTHU

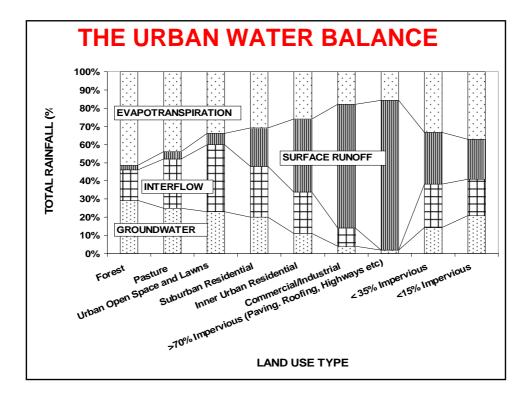
S. Barraud, Lyon 1 University / INSA Lyon &

A. Foulquier, Lyon 1 University, France









## ET IN THE URBAN WATER BALANCE

• Can be strongly affected by local conditions e.g climate, soil, topography etc. In arid regions for example, ET under a development scenario can be at least 80% - 97% of pre-development water balance.

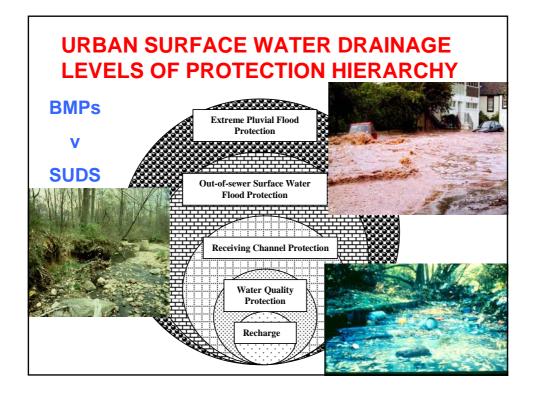
• Application of deep infiltration techniques will almost certainly lead to infiltration levels greater than which "naturally" occur. Widespread recharge could exacerbate already high WT levels resulting in changes to local habitat (ephemeral streams to perennial; lengthening flow durations; changes in riparian vegetation) as well as subsoil geotechnical changes.

• ET is function of area available for ET along with storage capacity. If pre- v post-ET areas are unequal, it could be difficult to make this up with storage, especially for "back-to-back" storm events.

• Matching pre-and post-ET under dense development (such as Smart Growth/LID) may NOT be a realistic objective, especially for rainfall patterns that limit the time of ET storage recovery between events.

• Dense urban development is likely to always have to manage excess runoff rather than rely on infiltration/recharge.





# FROM ROOFTOP TO RECEIVING WATER

- Better site design; enhanced LID, Smart Growth, retrofit etc

- IC minimisation (??? Could lead to urban sprawl); downspout disconnection and small scale infiltration

- Rooftop harvesting and treatment

- Site bioretention (rain gardens, street planters, soft street landscaping)

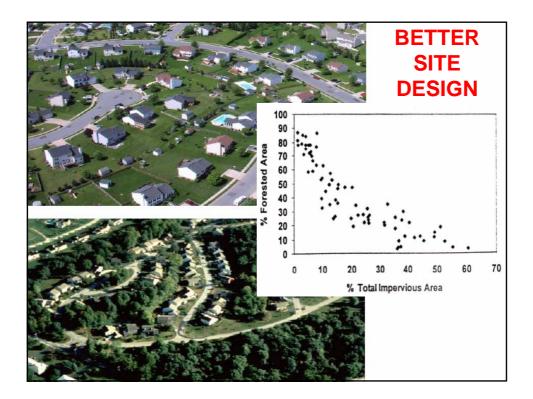
- Enhanced swale channels

- Sub-catchment BMPs (wetlands, retention/detention basins, infiltration basins)

- "Daylight" culverts

- Reforest streamside and "green" corridors -Eliminate illicit connections

- Target pollution "hotspots" and better ind/comm housekeeping
  - Increased public awareness and env campaigns; signage campaigns
  - Financial incentives and discounts for water saving storing schemes
    - Enhanced organisational and administrative arrangements/support for wider stakeholder consultation









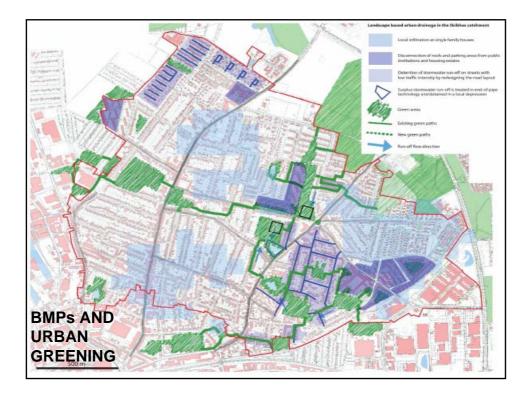












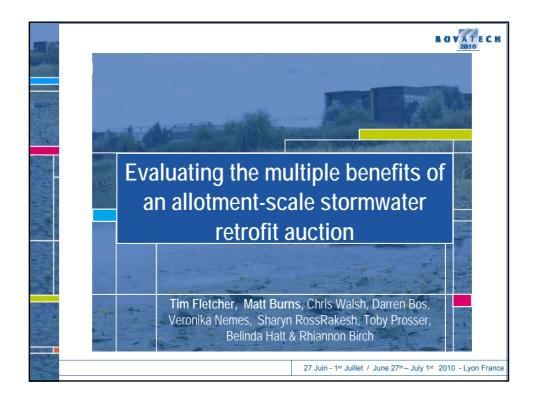


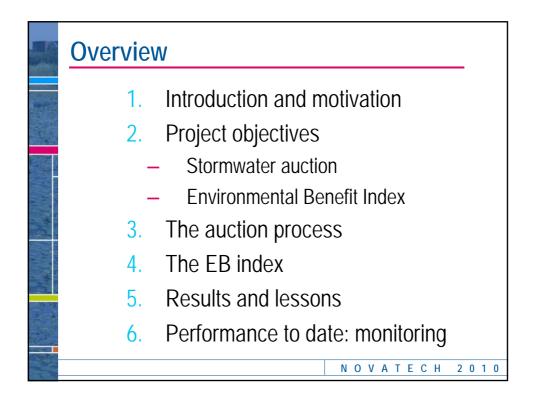
# 3<sup>rd</sup> GENERATION STORMWATER DRAINAGE Multi-layered, top-down and bottom-up approaches; address full range of flow events Jointly and concurrently considers plot, site and sub-catchment infrastructure design Enhanced LID basis (with modelling) and retrofitting Focussed on vegetative "leaf-out" analysis; 20%-30% minimum vegetation cover and reduction/suppression of IC effects Collaborative stakeholder agreement and public acceptance/awareness; fiscal support; supporting institutional frameworks and processes Regular maintenance schedules

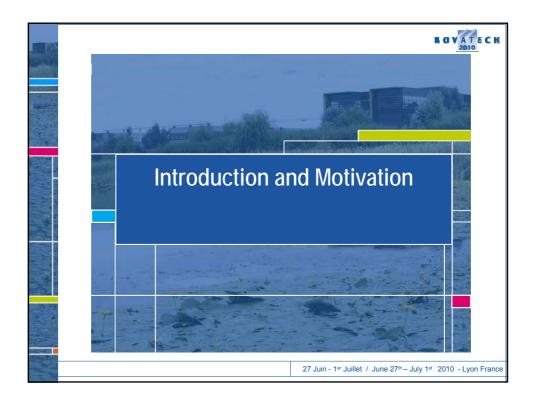
# LESSONS FROM A CATCHMENT-SCALE PUBLIC & PRIVATE-LAND RETROFIT PROJECT

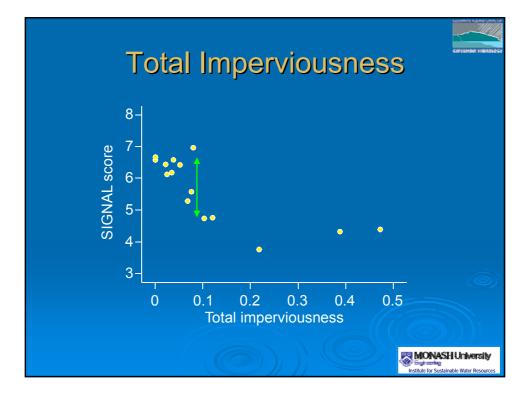
T. Fletcher & M. Burns

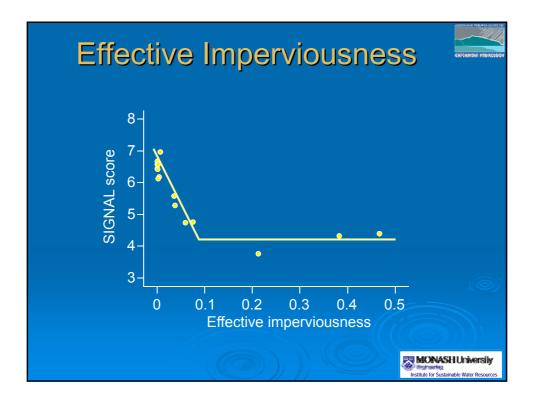
Monash University, Australia

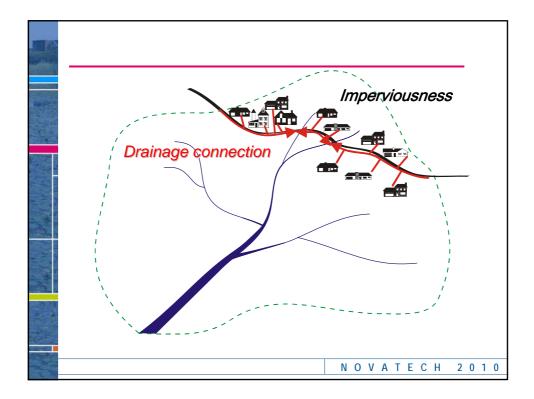


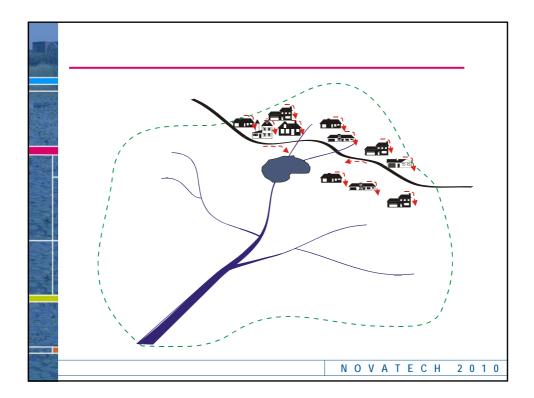




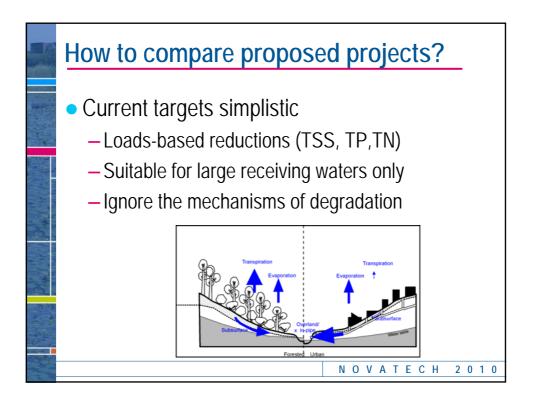


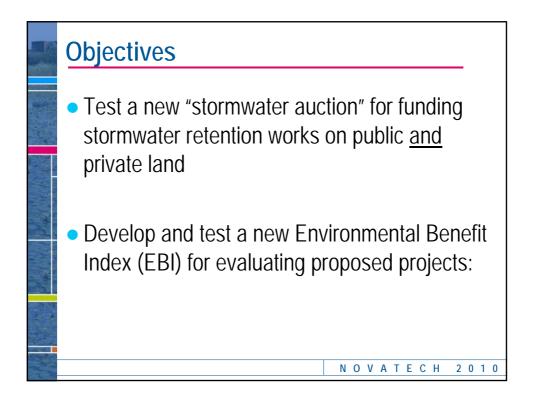


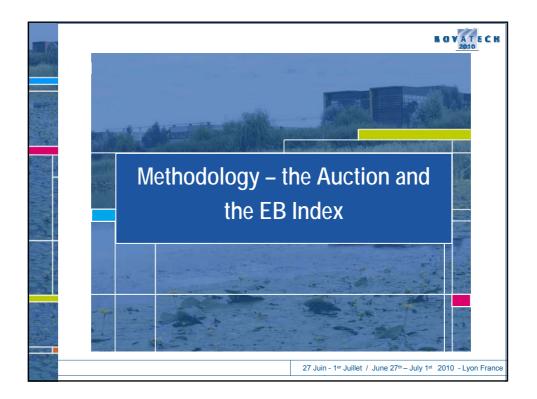


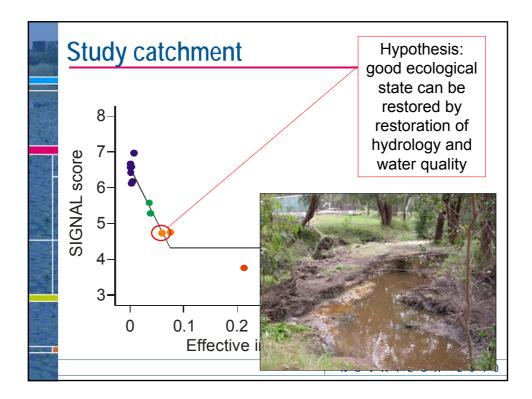


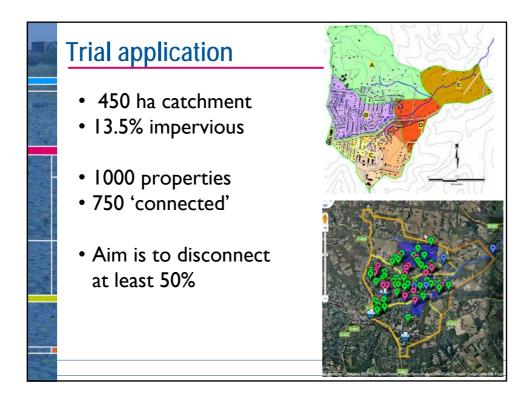


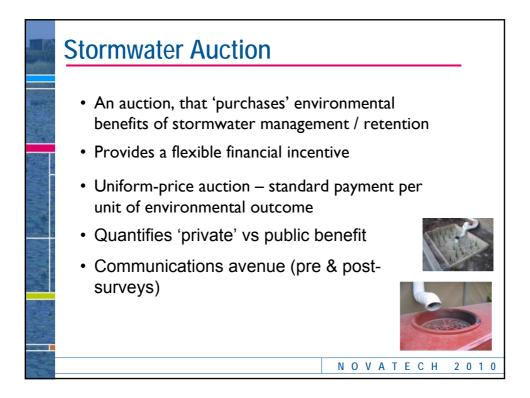






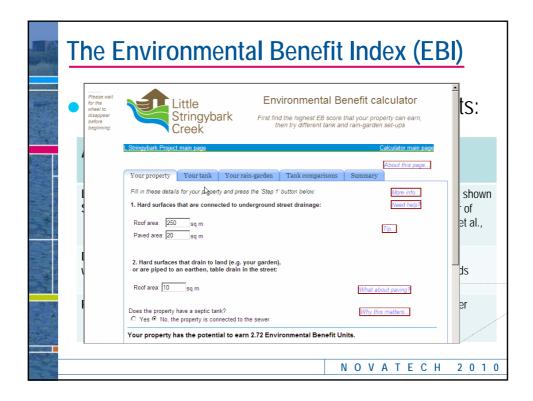


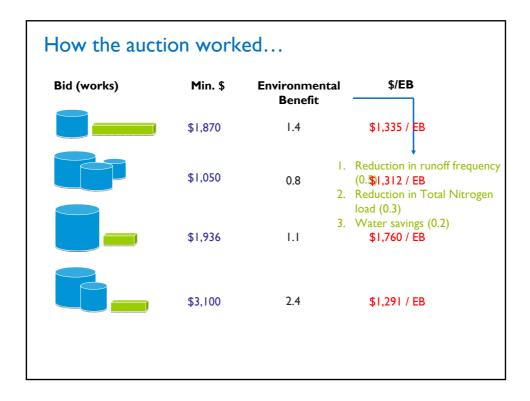


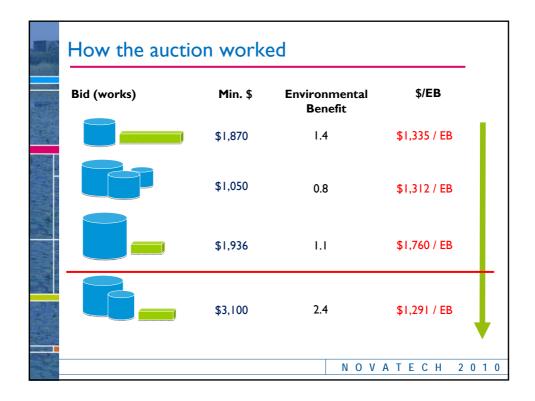


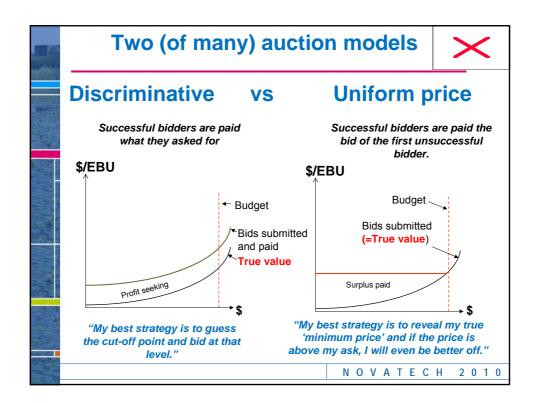




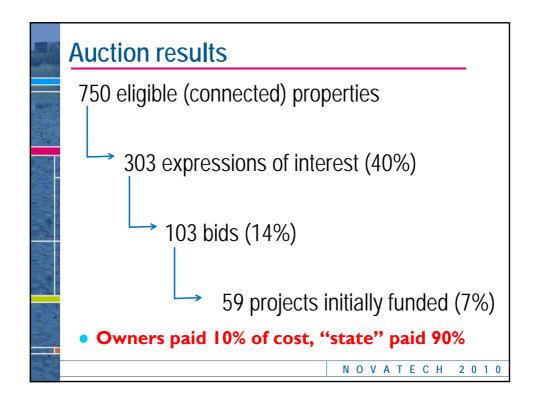




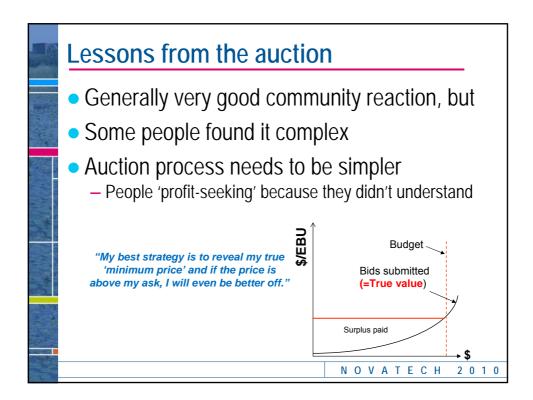


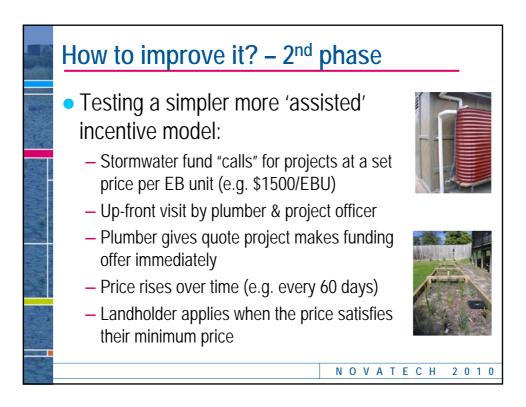


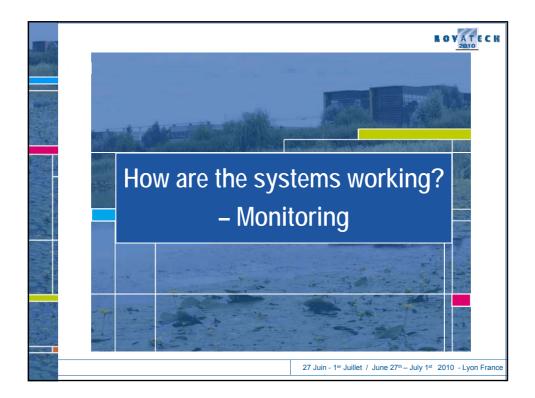


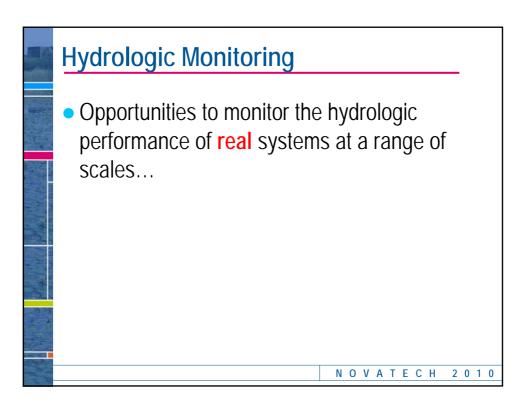


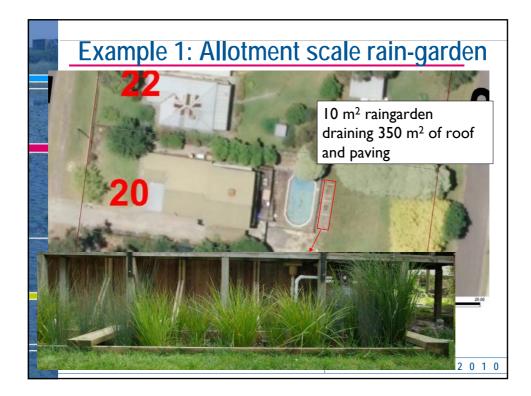


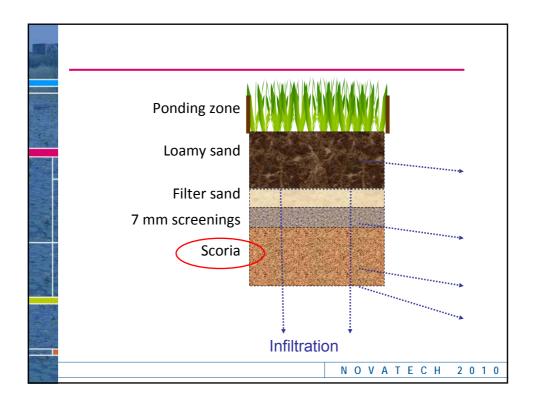


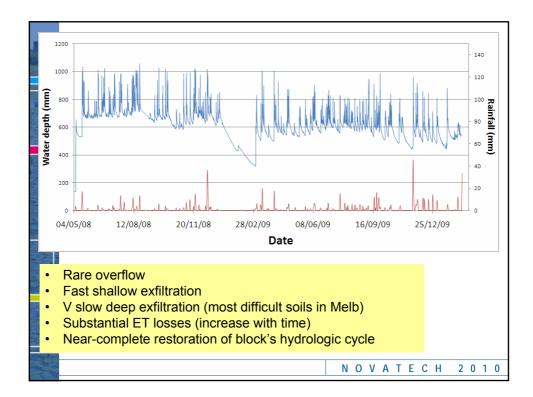


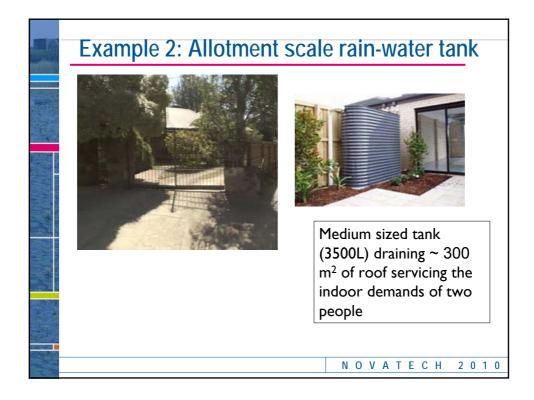


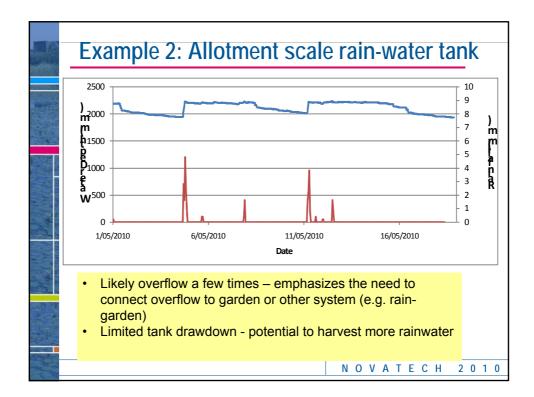


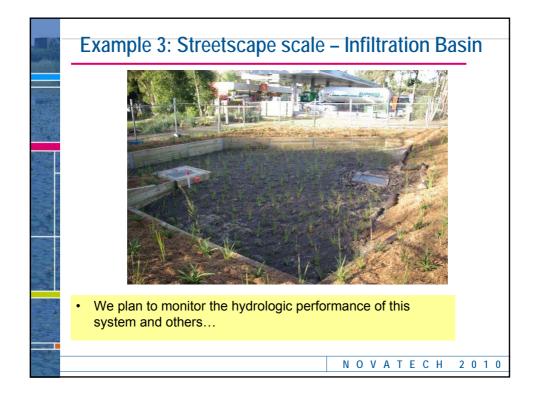


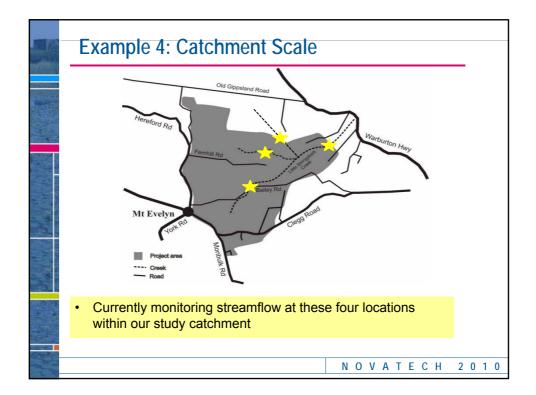








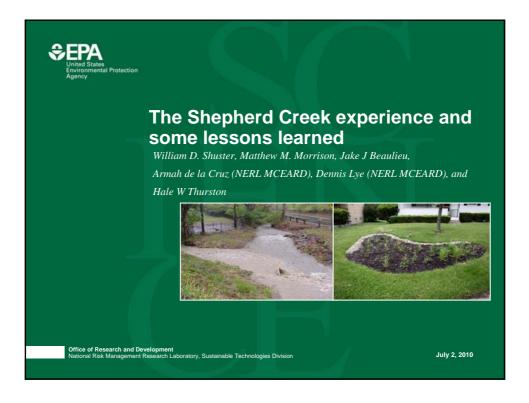


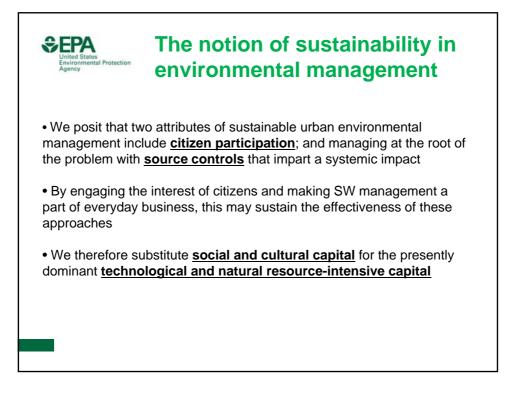


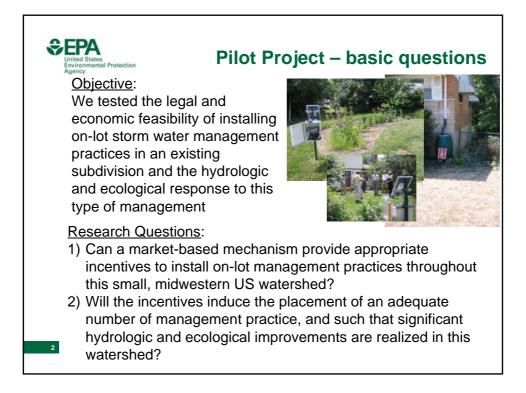
# LESSONS FROM THE SHEPHERD CREEK

B. Shuster

National Risk Management Research Laboratory, Office of Research and Development, USEPA, USA

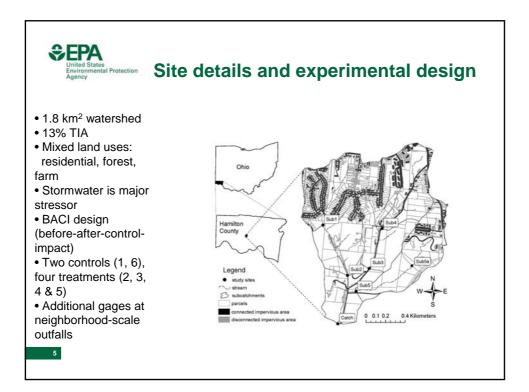


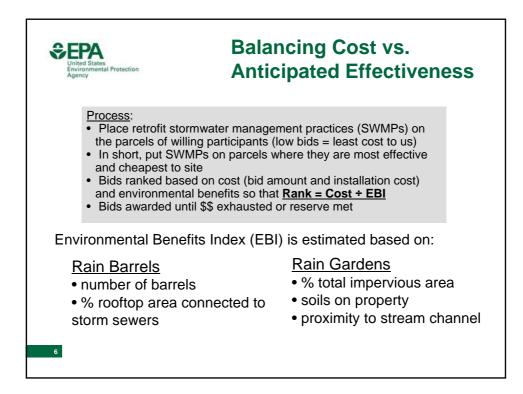


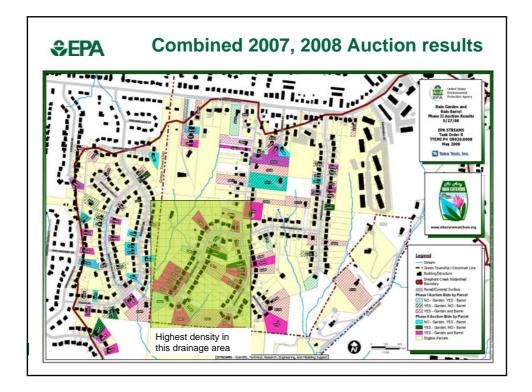


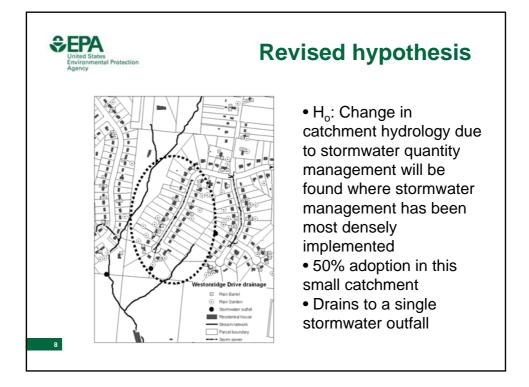


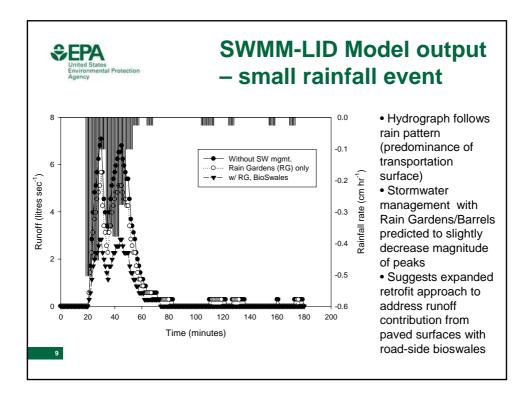


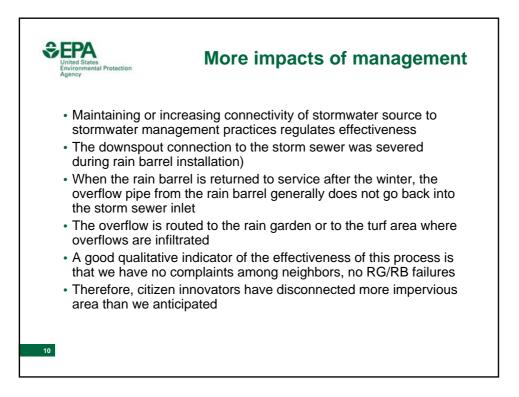


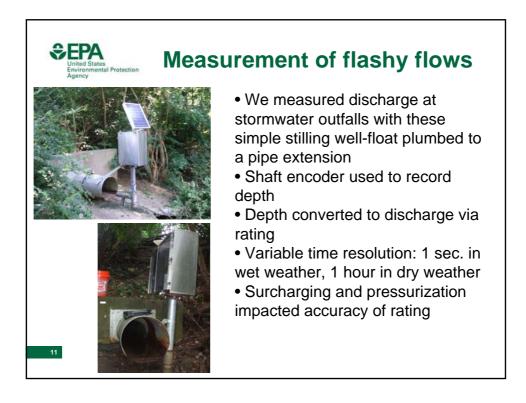


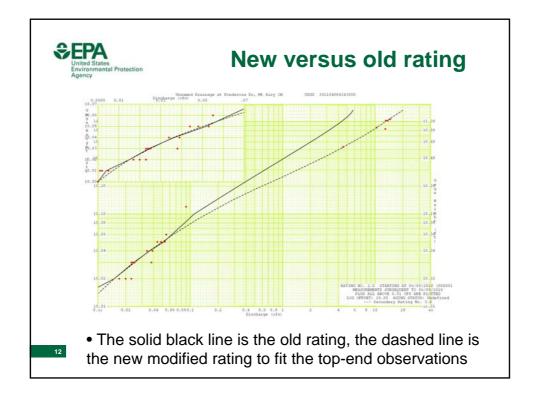


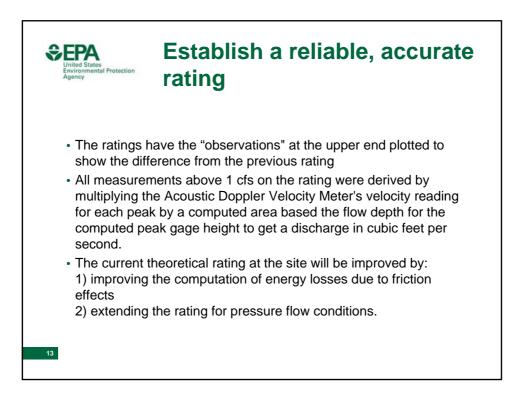


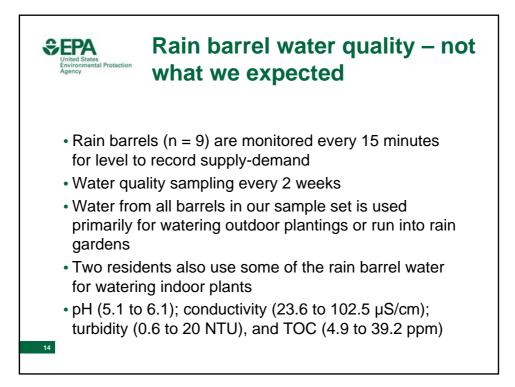


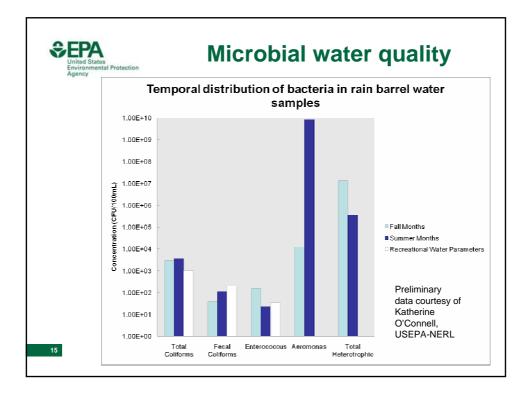


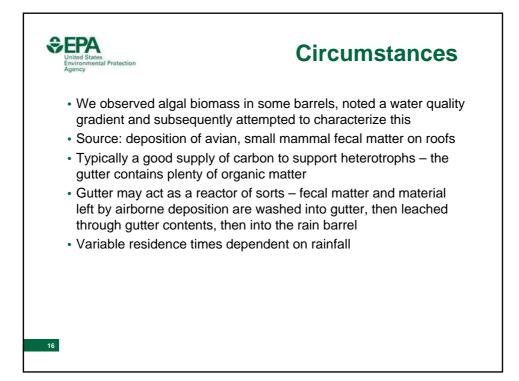


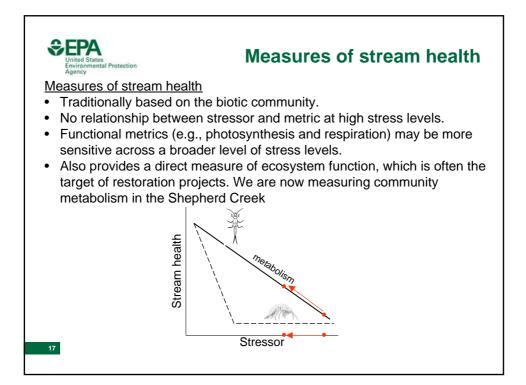


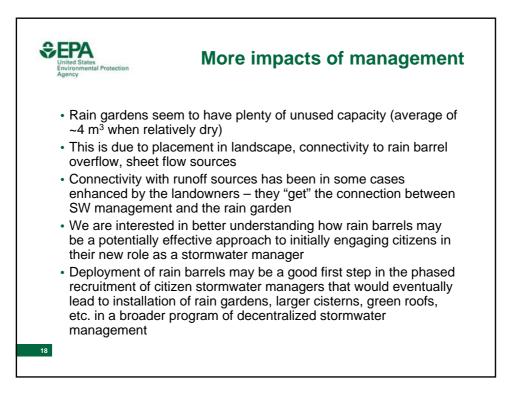


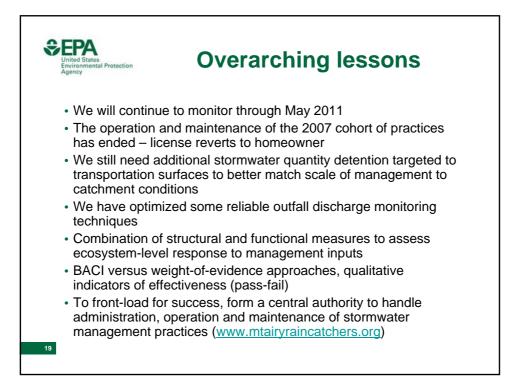








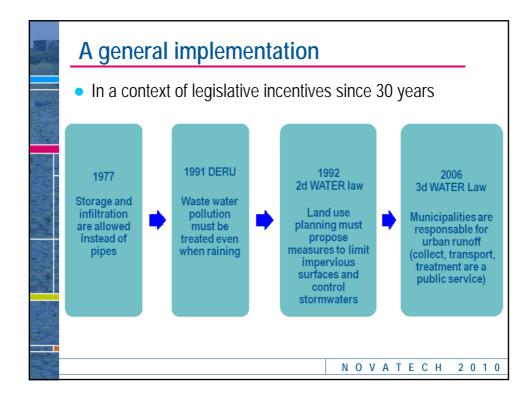


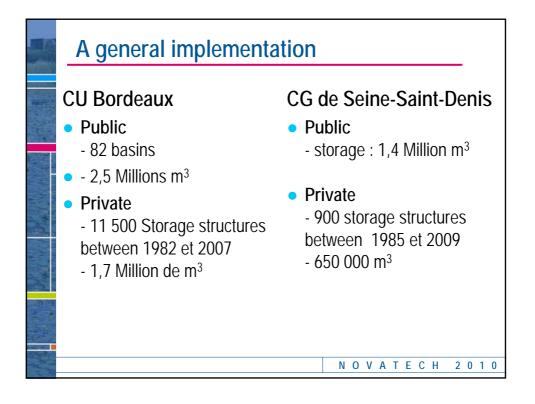


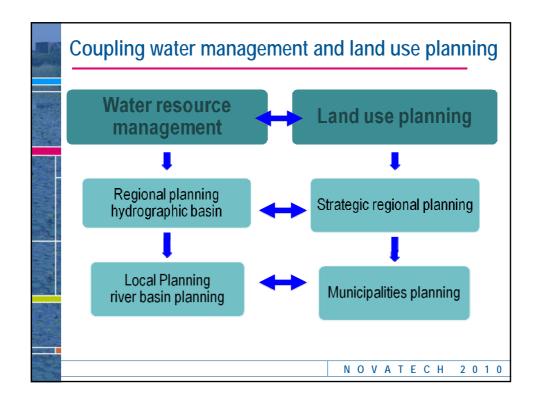
# SOME LESSONS LEARNT ABOUT SOURCE CONTROL STRATEGIES IN FRANCE

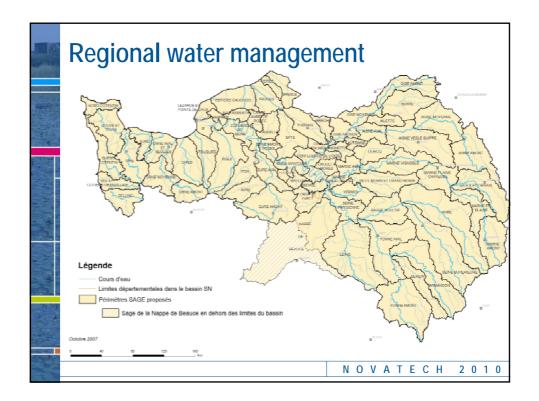
C. Carré Paris 1 University, France

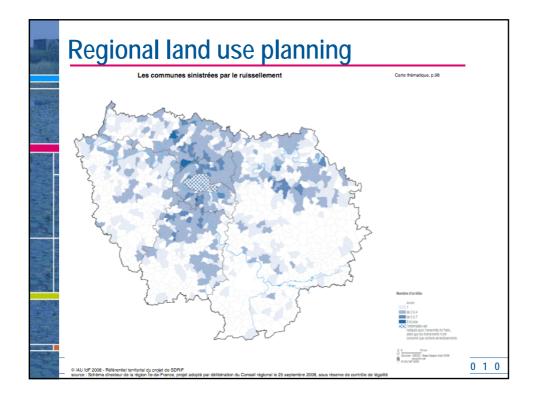


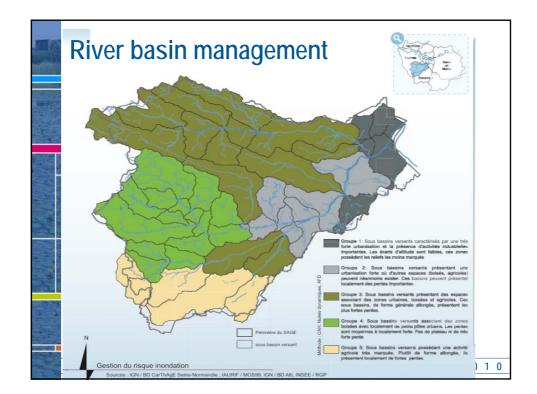


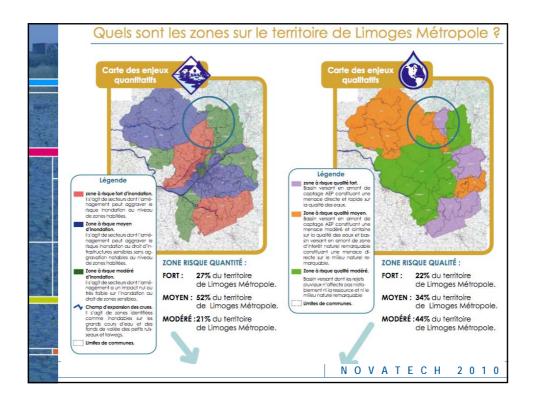




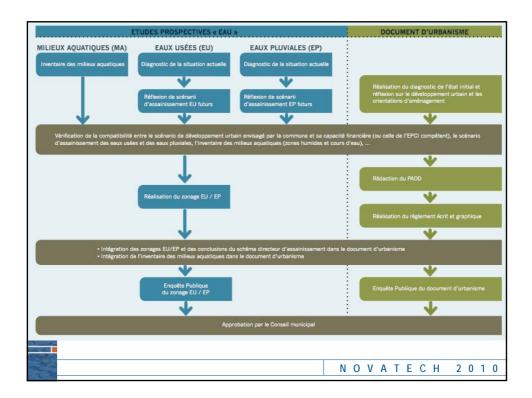




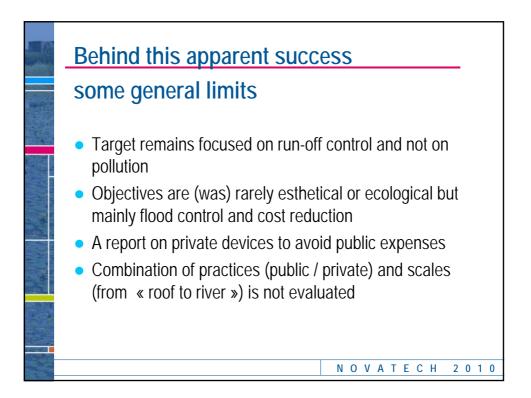


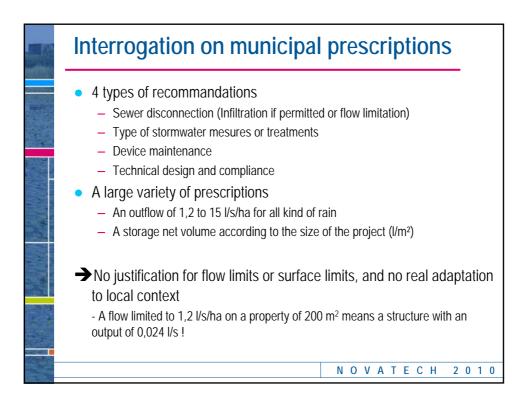


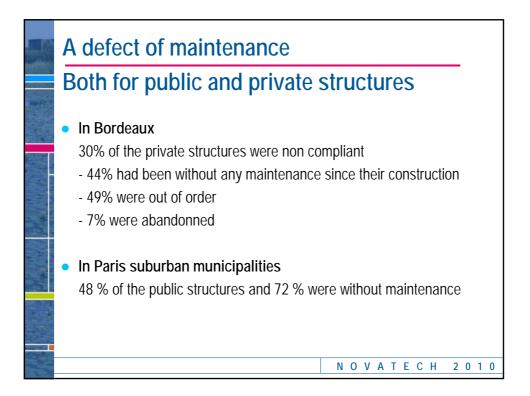




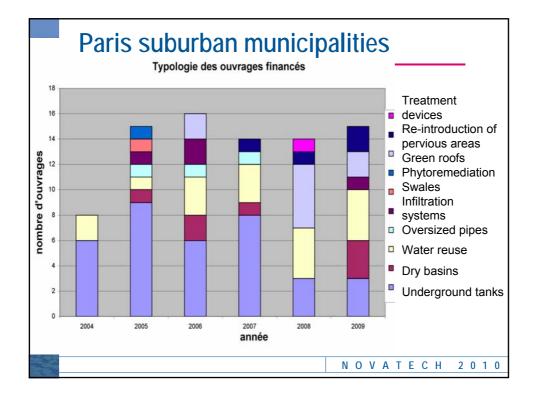




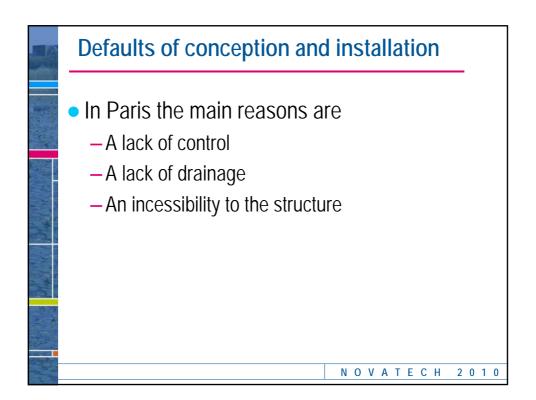




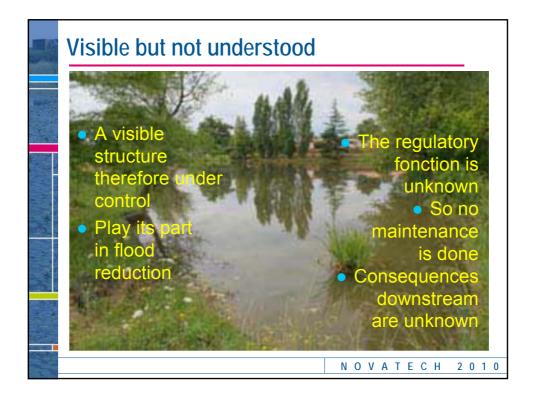
	Bordeaux private structures in 2007												
	Dry basins	Ponds	Storage pipe	Tanks	Swales	ro	een of / oof rage	Porous pavements	Source Infiltration systems	TOTAL			
	31	2	26	11	14		7	136	36	263			
	12 %	1%	10%	4%	5%	2	%	52%	14%	100%			
the second se	Par type de solution compensatoire												
· · ·	Bassin sec	Bassir eau	_			Noues	Toitu	re Structu		Total			
	31	2	26	11	l	14	7	136	36	263			
	11,8 %	0,7 %	10 %	4,2	% 5	5,3 %	2,6	% 51,7 9	% 13,7 %	100 %			
and the second s									•				
A A								NOVA	ГЕСН 2	2010			

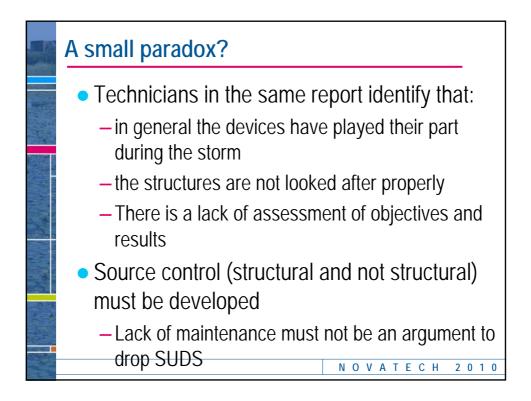


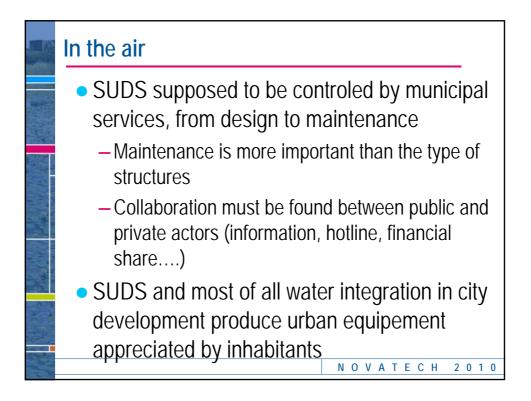
Propriétaire	Technique utilisée	Entretien réalisé	Fonctionnement correct
Collectivité	Bassin à ciel ouvert (87 m <sup>3</sup> )	Oui	Oui
Collectivité	Noue	Oui	Oui
Collectivité	Tranchée drainante	Oui	Oui
Collectivité	Puits d'infiltration	Oui	Oui
Collectivité	Canalisation surdimensionnée (814 m <sup>3</sup> )	Oui	Oui
Collectivité	Bassin à ciel ouvert (250 m <sup>3</sup> )	Oui	Oui
Collectivité	Bassin enterre (400m3)	Oui	Oui
Collectivité	Canalisation surdimensionnée (188 m <sup>3</sup> )	Oui	Oui
Collectivité	Bassin enterré (100 m <sup>3</sup> )	Oui	Oui
Collectivité	Bassin enterré	Oui	Non (pas de possibilité de vidange)
Collectivité	Puits d'infiltration	Non	Oui
Collectivité	Toitures terrasses minérales et végétales (110	Non	Oui
Collectivité	Bassin enterré (39 m <sup>3</sup> )	Non	Oui
Collectivité	Toiture terrasse minérale	Non	Oui
Collectivité	Toiture terrasse minérale (1900 m <sup>2</sup> )	Non	Non (pas de régulateur)
Collectivité	Canalisation surdimensionnée (50 m <sup>3</sup> )	Non	Non (pas de régulateur)
Collectivité	Canalisations surdimensionnées (368 m <sup>2</sup> )	Non	? (pas d'accès au bassin et au régulater
Collectivité	Bassin enterré (60 m <sup>3</sup> )	Non	? (Pas d'accès au bassin)
Collectivité	Toiture terrasse végétale	Non	Oui
Privé	Toiture terrasse minérale	Oui	Oui
Privé	SAUL (30 m <sup>2</sup> )	Oui	Oui
Privè	Bassin enterré (20 m <sup>2</sup> )	Oui	Oui
Privé	Puits d'infiltration	Oui	Oui
Privė	Toiture terrasse minérale (1700 m <sup>2</sup> )	Oui	Oui
Privė	Bassin enterré (18 m <sup>2</sup> )	Non	Oui
Privé	Canalisation surdimensionnée (1200 m <sup>3</sup> )	Non	Oui
Privé	SAUL (43 m <sup>3</sup> )	Non	Non (régulateur non entretenu)
Privé	Canalisation surdimensionnée (25 m <sup>3</sup> )	Non	Non
Privé	Bassin enterré (32 m <sup>3</sup> )	Non	? (Pas d'accès)
Privé	Bassin enterré (88 m <sup>3</sup> )	Non	? (Pas d'accès)
Privé	Bassin enterré (42 m <sup>3</sup> )	Non	? (Pas d'accès)
Privé	Bassin enterré (31 m <sup>2</sup> )	Non	? (Pas d'accès)











# LESSONS LEARNT AND EXPERIENCES ABOUT SOURCE CONTROL STRATEGIES IN BRASIL

N. Nascimento

UFMG, Brazil

# Lessons learnt and experiences about source control strategies in Brazil

SOCOMA Workshop on Design, modelling and implementation of source control technologies

> Nilo Nascimento Lyon, 27<sup>th</sup> June 2010

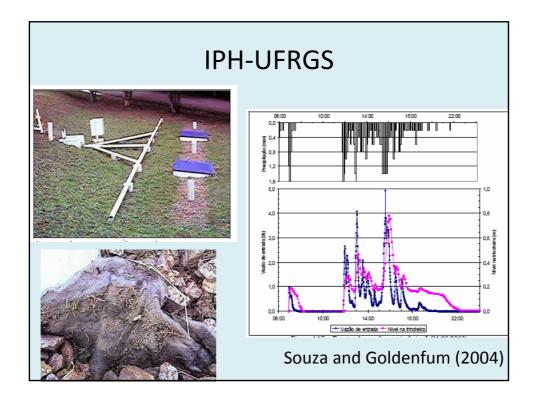


# Lessons learnt and experiences about source control strategies in Brazil

Introduction:

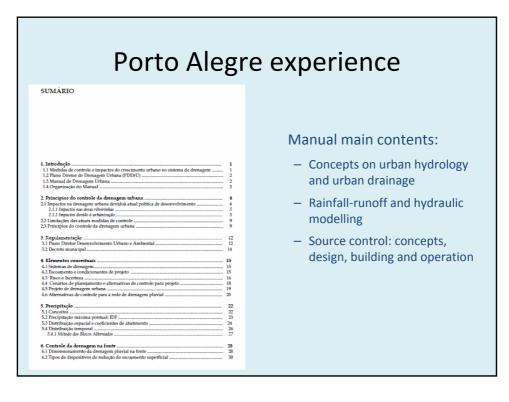
- Urban drainage for a long time influenced by traditional concepts
- Process of innovation following the democratization process (from the 1990'):
  - Participatory process: city councils, participatory budgeting
  - Close cooperation between municipalities and universities
- Two examples: Porto Alegre and Belo Horizonte







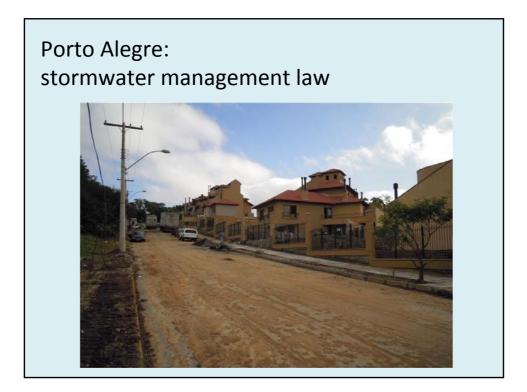
Porto Alegre experience							
Prefeitura Municipal de Porto Alegre DECRETO N ° 15.371 de 17 de novembro de 2006	PREFEITURA MUNICIPAL DE PORTO ALEGRE DEP - DEPARTAMENTO DE ESGOTOS PLUVIAIS						
	PLANO DIRETOR DE DRENAGEM URBANA Manual de Drenagem Urbana						
	Volume VI						
	Instituto de Pesquisas Hidráulicas Univerzidade Federal do Rio Grande do Sul						
	Setembro/2005						



# Porto Alegre: stormwater management law New urban developments are limited to a specific flow of 20.8 l/s.ha to drain to the public sewer system In order to meet this requirement, urban developers may employ storage devices to be designed as follow: If A < 100 ha, Storage volume = 4.25 l<sub>surf</sub> If A > 100 ha, hydrologic design for T = 10 y

### Porto Alegre: stormwater management law

- Urban developers may also employ other source control alternatives:
  - Draining to connected pervious areas: 40% reduction on I<sub>surf</sub>
  - Draining to impervious pavement: 50% reduction on I<sub>surf</sub>
  - Draining to non-connected perv. areas: 80% reduction on I<sub>surf</sub>
  - Draining to infiltration trenches: 80% reduction on I<sub>surf</sub>







## Porto Alegre: environmental law: compensating measures

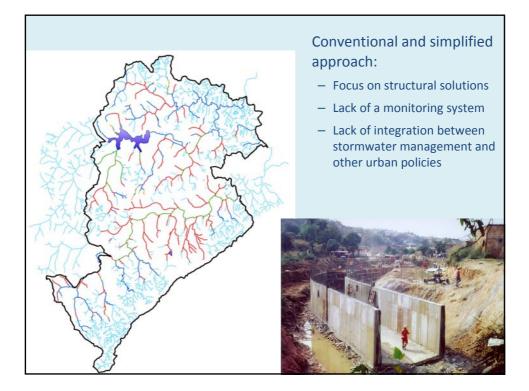


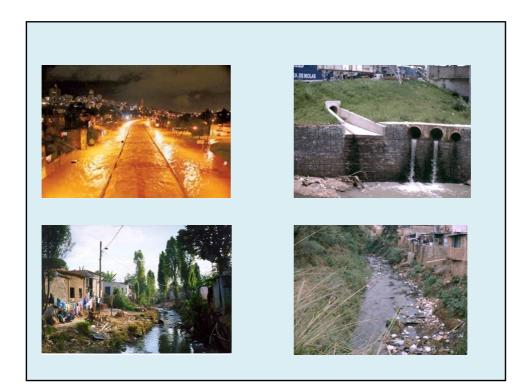


### Porto Alegre: stormwater management law

- Some results of this policy (source: Tucci ,2009):
  - 40 new urban developments adopted source control measures
  - Most of them are detention structures: dry DB set up in green areas
  - 30% are privately operated (land owners)
  - 70% are operated by the municipality

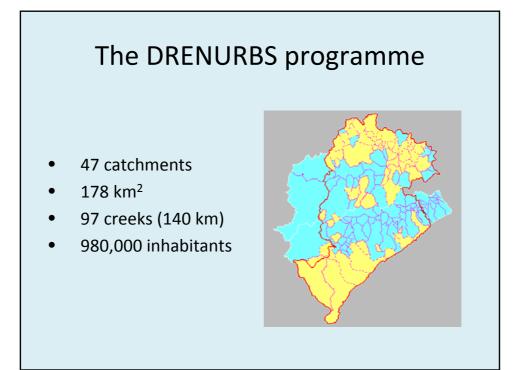


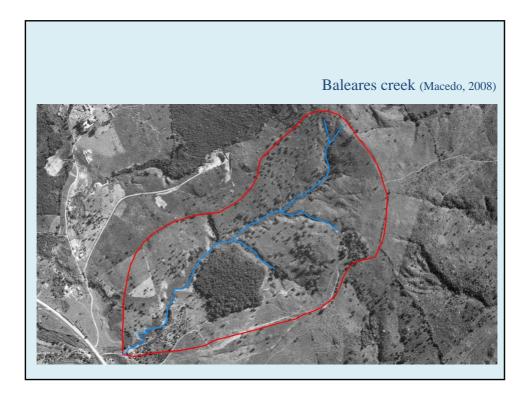


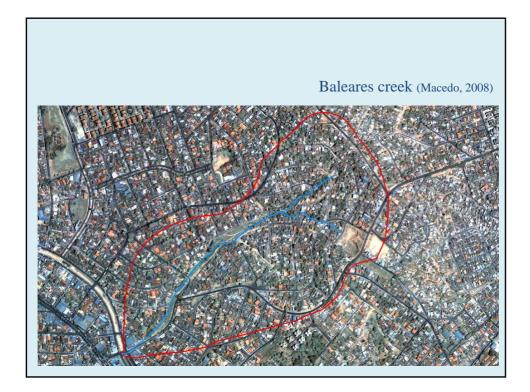


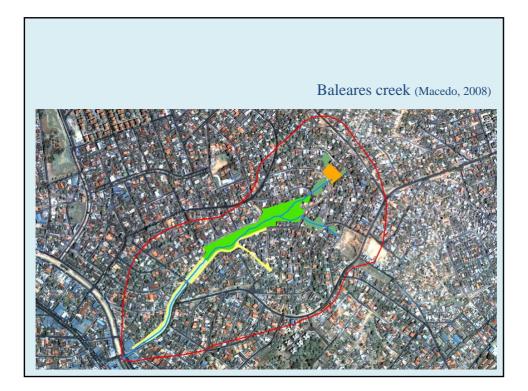


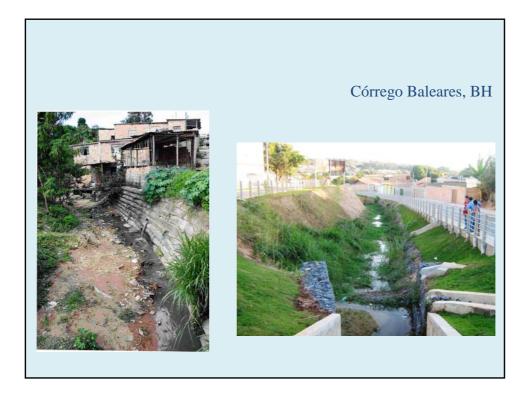
# <section-header>The DRENURBS programmeImage: Strain Str

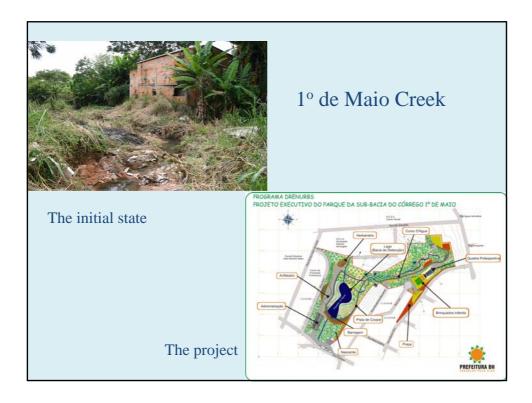


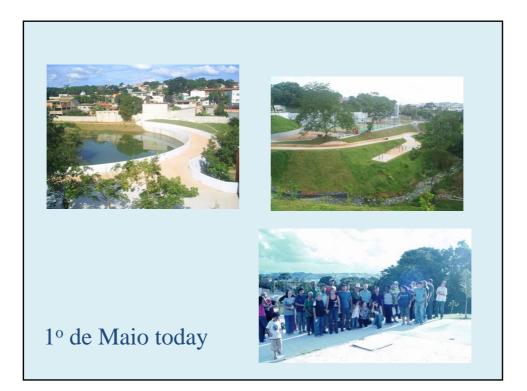




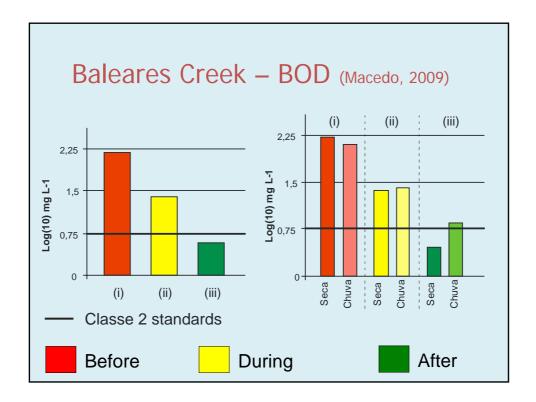


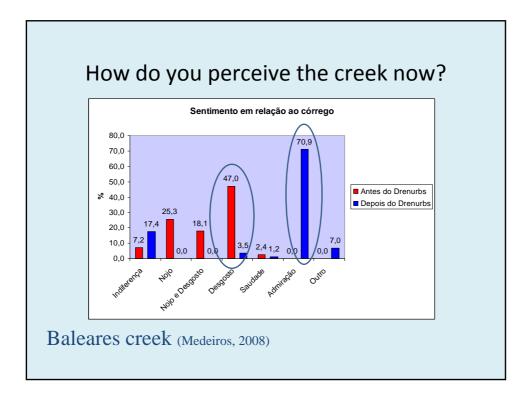


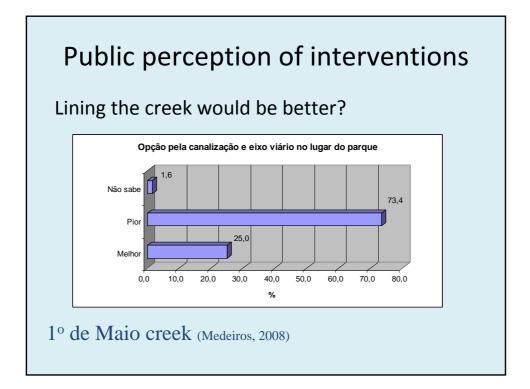


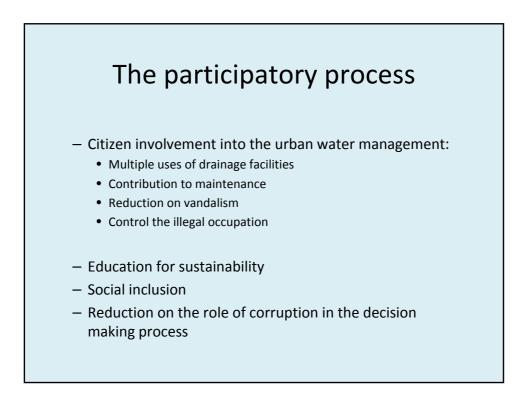


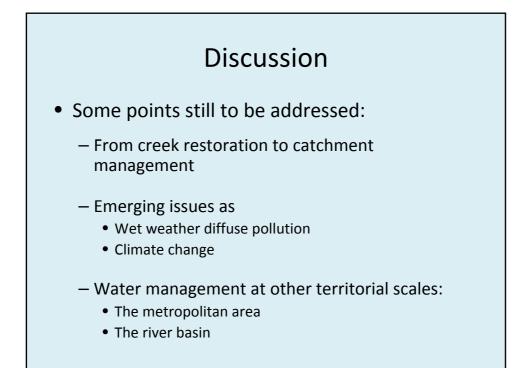


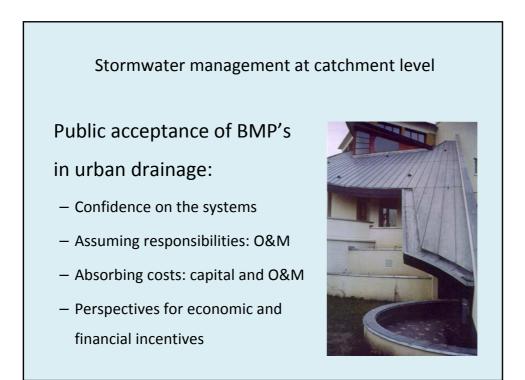


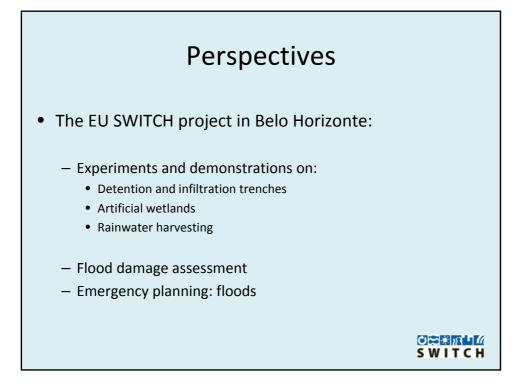


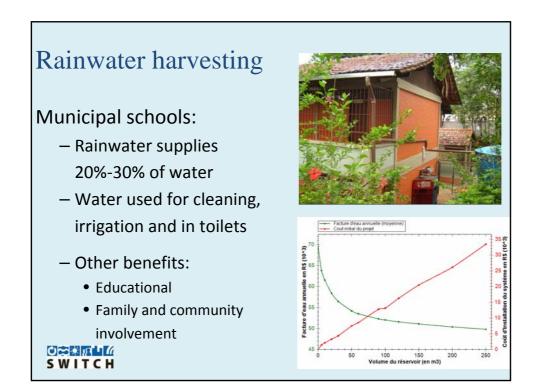


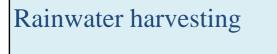












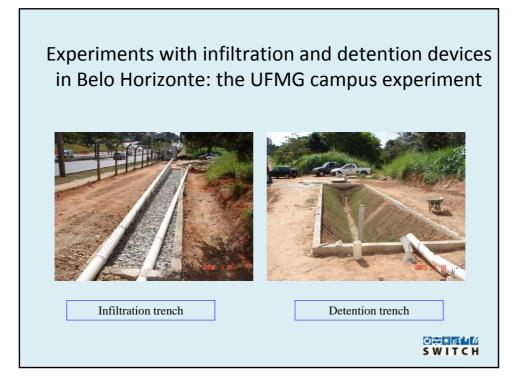
## Urban agriculture experiment:

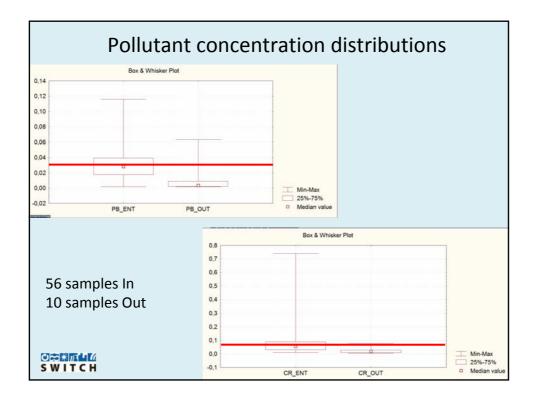
- Rainwater irrigation:
  - Supplies up to 40% of water
  - Overflows are infiltrated
  - Reduction on runoff and WWDP
- Other benefits:
  - Recycling organic wastes
  - Food security and income
  - Cultural memory and solidarity

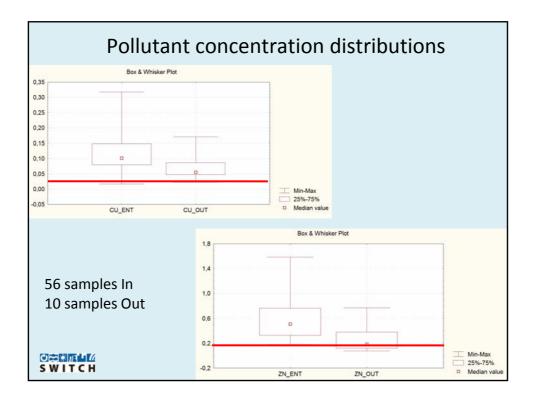
SWITCH

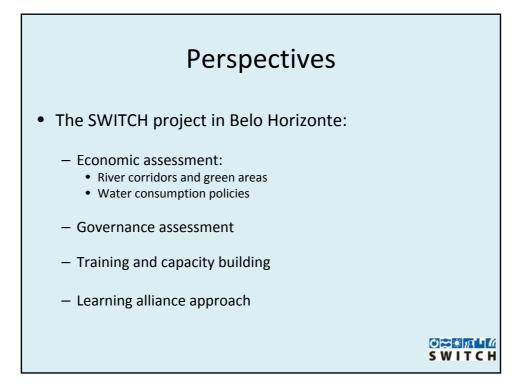


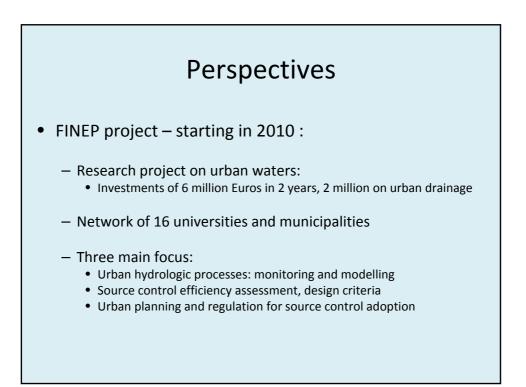














## **FORUM / DISCUSSION**

Coordinated by:

G. Rivard



